## **Final Report**

# <span id="page-0-0"></span>**Evaluation of Portable Emissions Measurement Systems (PEMS) for Inventory Purposes and the Not-To-Exceed Heavy-Duty Diesel Engine Regulation**

**Contract No. 03-345** 

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**Prepared for:** 

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#### **Disclaimer**

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#### **Executive Summary**

<span id="page-8-0"></span>Diesel engines are significant contributors of emissions in air basins and diesel emissions have become increasingly scrutinized. This scrutiny has resulted in more regulation of diesel emissions and a desire to learn more about the actual in-use emissions from these engines rather than values measured with engine and chassis dynamometers. However, measuring on-board, in-use emissions is problematic, especially if current EPA reference methods must be used. This project was launched with a goal of measuring the performance of alternatives to EPA reference methods. Toward that end, the practical alternative to fixed laboratory measurements with EPA reference methods is the use of portable emissions measurement systems (PEMS) which became the focus of this research. The intent of this project was to test the PEMS over a wide range of engine operating conditions so the results would provide insight for PEMS use for emission inventories and for regulatory applications, especially compliance with the Not-To-Exceed (NTE) regulations. The specific project deliverables included quantifying both the accuracy and precision of commercially available PEMS relative to the federal reference methods (FRM) in the University of California, Riverside's (UCR's) mobile emissions laboratory (MEL).

#### **Emission Measurement Tests –Common Elements**

The exact emission measurements were divided into two tests. First we measured emissions from a backup generator (BUG) representing a stationary source and true steady-state engine operation. Second, we measured emissions from a heavy, heavy-duty diesel truck (HHDDT), representing a mobile source at quasi-steady-state and transient engine operation. Common elements in both series of tests included a test matrix that was designed to represent a broad range of operating conditions and was accepted by the stakeholder group before testing was initiated. For the actual engine operation and testing, all PEMS and MEL were simultaneously connected to minimize errors due to sample variation. Furthermore, the PEMS manufacturers provided staff, instruments and technical expertise so the instruments were operated as designed.

Another common test element was that all measured values from the FRMs were blind to the PEMS manufacturers until they reported their data. After the analyses, we reported the deviations in concentration and flow rate between the PEMS and the FRMs, the two independent and critical values needed for calculating emission rates.

#### **PEMS Units Tested**

At the beginning of the research program, ARB solicited participation from a number of PEMS manufacturers. As a consequence, a total of seven PEMS manufacturers were represented in this research program and as one manufacturer provided two instruments, a total of eight instruments were tested. Three instruments measured solely particulate matter (PM), three measured solely gaseous emissions and the last two units measured both PM and gaseous emissions. Several points need to be made about the test program.

First, the highest priority was establishing data for the gaseous instruments, and second, not all the PEMS systems were used for both the stationary and the mobile source units. The PEMS tested are listed below in alphabetic order and throughout the text are identified using a specific number that is included in the table.

The Horiba OBS-2200 unit (PEMS4b) used for the chassis dynamometer test program at the ARB laboratory in Los Angeles was an early prototype instrument with unproven software at the time it was provided for the test program. Horiba submitted this unit in response to strong urging from both EPA and ARB even though the product development team had not had the opportunity to test the prototype with a vehicle. During the week of testing several software related problems were discovered, which required on-site software revisions. Consequently, the early runs in this program produced no viable data because of software bugs. The later data presented in this report is still prototype quality with incorrect software compensations applied. Results are preliminary to finishing the product development and should not be considered representative of the production model performance. Additionally, PEMS6 for measuring PM was included in the study for the chassis dynamometer testing. After the testing, it was determined that the instrument had not been operating within manufacturers' specifications, so these data were not reported.



#### **Emission Measurement Tests –Stationary Source/Backup Generator**

In the first phase, flow rates and emission measurements from a backup generator were measured with four PEMS and compared with the values from the FRMs. The diesel engine was a CAT 3406C, similar in size to the HHDDT engine for the second phase and the selected loads ranged from 5% to 100% to allow a comparison of concentrations representative of levels found over a wide operating range of a diesel engine.

Comparative results are shown in Figure ES-1 for the critical gases,  $NO<sub>x</sub>$  and  $CO<sub>2</sub>$ . Note for  $NO<sub>x</sub>$ , and  $CO<sub>2</sub>$ , the deviation range depended strongly on the PEMS and the load. For  $NO<sub>x</sub>$ , PEMS #2 and #4 showed the best agreement with the FRM, ranging from about 1 to 10% of the FRM. Reasons for observed deviations with the other PEMS are discussed in the report.



**Figure ES-1. PEMS Emission Rates Relative to FRM. Actual Emissions (g/hp-hr) Overlaid for Reference.** 

Comparative results for PM and THC are shown in Figure ES-2 for the BUG testing. Overall, the PEMS also showed relatively large differences relative to the FRM for THC. It should be noted that the absolute THC values overall are low, even in comparisons with the upcoming proposed NTE standards (see section 1.1 for a discussion of NTE control areas). The PM measurements for PEMS1 were considerably lower than those of the FRM while those for PEMS 3 were  $\sim$ 25% lower than the FRM at the high low points. The larger differences at the lower flow rates for PEMS3 can be attributed to a problem with flow measurement.



 **Figure ES-2 PEMS Emission Rates (g/hr) Relative to FRM at Several Loads.** 

## **Emission Measurement Tests –Mobile Source/ Heavy-duty Diesel Truck**

In the second phase of testing, emissions were measured from truck with a 475 hp Caterpillar C-15 ACERT engine certified to the 2.5 g/bhp-hr  $NO_x + NMHC$  and 0.1 g/bhp-hr PM standards. The testing compared emission measurements from seven PEMS with values from the FRM. Three of the PEMS measured only PM, three PEMS measured only gases and the remaining PEMS measured both gas and PM. The ACERT engine/vehicle system was tested on the ARB's heavy-duty chassis dynamometer.

Testing for this phase received much attention because results were likely to aid in the ongoing discussions about the measurement allowance for EPA's new regulation on inuse testing of diesel engines. The new regulation covers measurement of emissions from in-use diesel engines, especially when the engine operated in the Not to Exceed (NTE) region of the engine map. Accordingly, the testing was carried out while following a carefully planned series of quasi steady state and transient driving schedules. The four driving schedules included:

- 1) Three-mode quasi steady-state NTE test cycle.
- 2) Stepped NTE test cycle
- 3) The Urban Dynamometer Driving Schedule (UDDS)
- 4) The 50-mph Cruise mode from the ARB 5-mode HHDDT test cycle

The steady-state NTE cycle was run at three different engine speeds with events in the NTE region. A second NTE cycle was designed with a "stepped" pattern, intended to simulate NTE-type vehicle/engine operation with gentle accelerations between modes and steady-state operation at each load point. The NTE driving schedules test cycles were designed to clearly delineate entry into, operation within, and exit from the NTE-defined zone of engine operation. The UDDS and 50-mph cruise cycle were included to better represent 'real-world' operation and data in current emission inventories.

While data were collected for all four driving schedules, the executive summary focuses on results for the 3-Mode steady state NTE cycle since the results from other cycles were generally comparable. The executive summary details results for solely  $NO<sub>x</sub>$ , the highest priority target during the first phase of NTE regulation. Results for the 3-mode steady state cycle were averaged over the NTE events for a particular speed/load point for both PEMS and the FRMs. Figure ES-2a provides the average NTE results in g/bhp-hr for  $NO<sub>x</sub>$  over the various modes and Figure ES-2b shows the differences relative to the NTE threshold value as the error margin relative to the standard is an important regulatory consideration. A total of 16 speed/load points are included in the Figures. Data collected at 1290 rpm, 40% load did not meet the criteria for an NTE event and not included in the Figures.

Again results depended on the PEMS. For PEMS2 the  $NO<sub>x</sub>$  emission rates were approximately 3 to 17% higher than those for the FRM in g/bhp-hr units relative to an NTE threshold of 2.0 g/bhp-hr. PEMS 4 showed larger differences compared to the FRM, with percent differences ranging from 26 to 47%. It should be noted that PEMS4 for the chassis dynamometer testing was a prototype. As explained in the report, an extensive data analysis of the emissions, engine operation and other factors was possible for only two of the four gaseous PEMS.



#### Figure ES-3. (a) Comparison of NO<sub>x</sub> Emission Rates and 95% Confidence Limits. (b) **Comparison of Percentage Differences in NO<sub>x</sub> Emission Rates for Different PEMS.**

The  $CO<sub>2</sub>$  emissions are important since  $CO<sub>2</sub>$  is often used in determining fuel specific (fs) emissions. Results in Figure ES-3 indicate that during steady state NTE events,  $CO<sub>2</sub>$ emissions for PEMS2 were generally about 5% higher than those of the FRM and PEMS4 values were 9 to 30% higher than those the FRM. Further analysis is offered in the main report.



Figure ES-4. Percentage Difference between PEMS and FRM for CO<sub>2</sub> in g/bhp-hr for NTE **Steady State Cycle** 

THC over the NTE steady state cycle was generally lower for both PEMS2 and PEMS4 compared with the FRM. THC emissions were below the NTE threshold standards for these measurements. Relative to the approximate NTE threshold for THC of 0.21 g/bhphr, the percentage differences were all within 10%, with most being within 5%.

CO emissions for PEMS2 showed good correlation with the FRM, but measurements were consistently higher than those for the FRM. Again, however, the CO emissions were well below the NTE threshold of 19.4 g/bhp-hr, and the percentage differences between the FRM and PEMS2 were all within 2% relative to the NTE CO threshold.

During the mobile source testing, an exploratory trial was made of four PM-capable PEMS. Results showed real promise for the new instruments to follow the modal behavior of the PM emissions. PEMS3 and PEMS7 showed the best correlation with values within 15-25% of the PM mass emissions measured by the FRM, with considerably larger deviations seem for some other instruments.

## **Findings & Recommendations**

The results of this research show that of the four gaseous PEMS that PEMS2 compared better with the FRM for  $NO<sub>x</sub>$  and  $CO<sub>2</sub>$  than other units. An exploratory trial of PMcapable PEMS show great promise for real-time PM measurements but more development is needed. The following recommendations are made based on the observations in this research:

- More attention is needed to establish detailed use protocols and performance limits as specified for a CFR method. For example, the CFR specifies a particular type of analyzer for  $NO<sub>x</sub>$  as well as the specifications.
- Even the best PEMS with an experienced operator had problems so PEMS require experienced operators.
- It is suggested that more frequent calibrations are needed to establish confidence intervals and ensure reliable operation of the PEMS during sampling in the field.
- For the best progress, measurement programs must be collaborative in design and analysis between the PEMS manufacturer, ARB, EPA, EMA and the contract lab making measurements.

Looking ahead, we expect PEMS confidence limits to improve from that observed in this research as a result of the ARB/EPA/EMA's Measurement Allowance Program.

## **Next Steps**

The original scope for this project included four main emissions testing tasks: 1) stationary source/backup generator testing of all participating PEMS, 2) mobile source/heavy-duty truck chassis dynamometer testing of all participating PEMS, 3) Onvehicle, over-the-road testing of all participating PEMS using pre-selected test routes, and 4) on-board, over-the-road testing of "the most suitable" PEMS under conditions meant to mimic the actual in-use compliance program (i.e., real vehicles in actual revenue service). However, since this project began, the ARB, US EPA and heavy-duty engine companies (as represented by the Engine Manufacturers Association) agreed to a measurement allowance program to determine exactly what numerical values should be

assigned, as referenced in the Memorandum of Agreement (MoA) signed by the above parties in June, 2005. The original scope of this project is being revised so that the resources originally allocated for this project can reallocated to support this measurement allowance program. The measurement allowance program will be examining topics and issues similar to those studied in this project - how PEMS compare against the reference methods, but will do so in a more focused and extensive manner. Specifically, all factors expected or suspected to influence PEMS emissions measurements such as vibration, ambient conditions (such as changes and variations in pressure, temperature. and humidity, ambient background HC concentrations), radio frequency and electromagnetic interference, etc. will be studied with the goal of developing actual, pollutant-specific, numerical measurement allowances for  $NO<sub>x</sub>$ , THC, and CO. In addition, the measurement allowance program will also quantify variability between engine dynamometer measurements of engine torque and speed. The current project scope is being revised to support this measurement allowance program. Specifically, CE-CERT will: 1) perform a Code of Federal Regulations Part 1065 audit of the CE-CERT mobile emissions laboratory, 2) perform side-by-side testing at SwRI, and 3) perform a validation of SwRI's Monte Carlo model used to develop the measurement allowance by collecting over-the-road emissions data using the MEL.

## **1.0 Background**

<span id="page-15-1"></span><span id="page-15-0"></span>On-road heavy-duty diesel (HDD) engines/vehicles are projected to be a significant contributor to the emissions inventories for oxides of nitrogen  $(NO<sub>x</sub>)$  and particulate matter (PM), even in 2010, because of their high emission rates and the long lifetimes of the vehicles. While the regulatory emission standards have become more stringent over time, the introduction of electronic controls allowed engines to operate with advanced timing and to generate "off-cycle" emissions, specifically higher  $NO<sub>x</sub>$  and lower PM levels than produced over the laboratory engine dynamometer certification test cycle. In the late 1990's, the US Environmental Protective Agency (EPA) and the California Air Resources Board (CARB) signed agreements with some engine companies that required, among other things, supplemental tests used for certification of heavy-duty diesel engines. The supplemental tests include the in-use Not-To-Exceed (NTE) test procedure, the EURO III European Stationary Cycle (ESC) test procedure, and measurement of emissions within the NTE control area.

Effective monitoring equipment is needed to assess in-use HDD emissions. This project evaluated current commercially available portable emissions measurement systems (PEMS). The project is intended to represent a snapshot of the current state-of-art and a reference point from which PEMS could be evaluated as a tool for either emissions inventory development or in-use compliance. This PEMS evaluation determined the basic measurement capabilities (e.g., accuracy, precision, etc.) and compare those results with measurements made with laboratory grade emissions analyzers that are specified in the federal reference methods (FRM) and as used in the University of California at Riverside's (UCR)'s mobile emissions laboratory (MEL). The goal of the project is to evaluate the PEMS suitability for inventory/model building work and for the NTE compliance work.

The PEMS systems selected for inclusion in the project were evaluated over two main test programs with increasing levels of complexity in measurement and potential variability in operation. In the first test, the PEMS were evaluated using emissions from a diesel engine driving a back-up generator (BUG) over a series of load points. In the second portion of the test program, the PEMS were compared over chassis dynamometer test cycles. The chassis dynamometer tests included transient cycles, steady state cycles, and cycles designed to provide test conditions that will create NTE events and the transition into and out of the NTE zone. The chassis dynamometer transient tests represent an additional level of complexity for the PEMS measurements in that they require accurate measurement and correlation of the exhaust concentrations and the exhaust flow rate. The chassis dynamometer testing was designed to operate the PEMS under conditions with the engine having more degrees of freedom or sources of error and closer to the NTE test conditions than the backup generator testing. Both tasks represent conditions that are more controlled than actual over-the-road operation. Additional work for comparisons of PEMS with the UCR MEL during over the road measurements is currently being planned as part of the ARB/EPA/EMA's Measurement Allowance Program.

### **1.1 Data Analysis with a Focus on the NTE Zone**

<span id="page-16-0"></span>The original intent of the project was to broadly survey the accuracy of the PEMS when applied to data gathering for either NTE or inventory purposes. However, as the project developed, a greater interest was assigned to defining the accuracy and percent deviation of the PEMS within the NTE zones. The NTE zones were defined by agreements between the US EPA, CARB and the engine manufacturers with more information provided in the EPA documents<sup>1</sup>. Paraphrasing the reference: An NTE event is generated when all of the following conditions are simultaneously met for at least 30 seconds or longer if an after treatment device is regenerating.

A listing of NTE conditions is provided in Table 1-1 and the NTE region is illustrated graphically in the Figure 1-1.

1. Speed > $15\%$ (n <sub>hi</sub> -n <sub>lo</sub> ) + n <sub>lo</sub>	7. Outside petitioned exclusion zones				
2. Torque $\geq$ 30% max	8. Outside of any NTE region in which a manufacturer states				
	that less than 5% of in-use time will be spent.				
3. Power $\geq$ 30% max	9. With EGR, intake manifold temperature $\geq 86{\text -}100^{\circ}F$ ,				
	depending upon intake manifold pressure.				
4. Altitude $\leq$ 5500 feet	10. With EGR engines, the engine coolant temperature $\geq$ 125-				
	140°F, depending on intake manifold pressure.				
5. Amb temp $\leq 100^{\circ}$ sea level to	Engine after treatment systems' temperature $\geq 250^{\circ}$ C. 11.				
86°F at 5500 feet	Only for $NOx$ and HC aftertreatment.				
BSFC $\leq$ 105% min, non- 6.					
automatic, non-manual transmission;					
essentially for series hybrids					

**Table 1-1. Specifications for Events Classified in the NTE Zone** 



**Figure 1-1. Graphical Examples of the NTE Control Area** 

## **2.0 Experimental Design and Procedures**

<span id="page-17-0"></span>The emission factors used to characterize diesel engines are often based on engine and chassis dynamometer data rather than real-world, on-board, in-use measurements. However, measuring on-board emissions is problematic, especially if current EPA reference methods must be used. This project was launched with a goal of measuring the performance of alternatives to EPA reference methods. Toward that end, the practical alternative to fixed laboratory measurements with EPA reference methods is the use of portable emissions measurement systems (PEMS) which became the focus of this research. The intent was to test the PEMS over a wide range of engine operating conditions so the results could provide insight for PEMS use for emission inventories and for regulatory applications, especially compliance with the Not-To-Exceed (NTE) regulations. The specific project deliverables included quantifying both the accuracy and precision of commercially available PEMS relative to the federal reference methods (FRM) in the University of California, Riverside's (UCR's) mobile emissions laboratory (MEL).

#### **2.1 Emission Measurement Tests –Common Elements**

The exact emission measurement program was divided into two major test sequences. First, measure emissions from a diesel engine in a backup generator (BUG) representing a stationary source, and second, measure emissions from a heavy, heavy-duty diesel truck (HHDDT), representing a mobile source. Common elements in both series of tests included a test matrix that was designed to represent a broad range of operating conditions and was accepted by the stakeholder group before testing was initiated. CARB diesel fuel was used for all testing. For the actual engine operation and testing, all PEMS and MEL were simultaneously connected to minimize errors due to sample variation. Furthermore, the PEMS manufacturers provided staff, instruments and technical expertise so the instruments were installed, calibrated (QC/QA) and operated as designed. There were two exceptions to the PEMS manufacturers' staff operating the instruments. During the HHDDT test, one gaseous analyzer was operated by ARB staff and the PM devices were operated by UCR staff.

Another common test element was that all testing started with a comparison between the FRM and PEMS for 1% audit and blended calibration gases. Next the comparative data was obtained for both the FRMs and PEMS while the engines operated over the planned driving schedules. In any case, as part of the analyses protocol, all measured values from the FRMs were blind to the PEMS manufacturers until they reported their data. After the analyses, we reported the deviations between the PEMS and the FRMs in concentration and flow rate, the two independent and critical values needed for emission factors.

#### **2.2 Measurement Method – CE-CERT's MEL with FRMs**

<span id="page-18-0"></span>A full description of UCR's MEL is available in the peer-reviewed literature and a synopsis of key points are provided in Appendix  $A<sup>2</sup>$  During all of the comparative testing, CE-CERT's MEL measured nitrogen oxides  $(NO<sub>x</sub>)$ , nitrogen dioxide  $(NO<sub>2</sub>)$ , total hydrocarbons (THC), non-methane hydrocarbons (NMHC), carbon monoxide (CO) and carbon dioxide  $(CO_2)$  and particulate matter (PM) emissions. The key experimental point is that all pollutants were measured using analyzers and the methodologies for calibration and operation that carefully followed the Code of Federal Regulations (CFR). These methods are referred in this report as Federal Reference Methods (FRMs). During each test cycle, both modal (second-by-second) and bag data for gaseous emissions were collected and analyzed. The bag data served as an independent check of the modal values as part of the quality assurance program. For PM mass emissions, the MEL collected a single, integrated sample for each test cycle run. A limited number of PM samples were collected on quartz filter in MEL to measure the elemental and organic (EC/OC) fraction of the PM emissions during the chassis dynamometer testing.

#### **2.3 Measurement Method -- PEMS Units Tested**

At the beginning of the research program, ARB solicited participation from a number of PEMS manufacturers. A total of seven PEMS manufacturers participated and were represented in this research program and as one manufacturer provided two instruments, a total of eight instruments were tested. A listing of the instruments is provided in Table 2-1, with a more detailed description of the instruments in Appendix B. Three instruments measured solely particulate matter (PM), three measured solely gaseous emissions and the last two units measured both PM and gaseous emissions. Several points need to be made. First, the highest priority was establishing data for the gaseous instruments, and second, not all the PEMS systems were used for both the stationary and the mobile source tests. The PEMS tested are listed below in alphabetic order and throughout the text are identified using a specific number that is included in the table. One additional PEMS for measuring PM was also initially included in the study for the chassis dynamometer testing. After completion of testing, however, it was determined that the instrument had not been operating with manufacturers specifications. These data are not included in the report.





As each PEMS manufacturer approached their instrument design in a unique manner, the measurement capabilities of each PEMS differed. Four of the systems had capabilities for measurement of gaseous emissions, with two these systems also capable of PM measurements. As the project evolved gaseous measurements became the highest priority so a summary of the measurement techniques for each of the PEMS capable of gaseous emissions measurements is provided in Table 2-2. The PEMS units for gaseous emissions are identified as PEMS1, 2, 3 and 4 throughout the report since the focus of the work was to provide a broad characterization of PEMS and their application for in-use measurements of diesel emissions, as opposed to an evaluation of specific technologies. The PEMS capable of only measuring PM are discussed in greater detailed in the results section below.

The Horiba OBS-2200 unit (PEMS4b) used for the chassis dynamometer test program at the ARB laboratory in Los Angeles was an early prototype instrument with unproven software at the time it was provided for the test program. Horiba submitted this unit in response to strong urging from both EPA and ARB even though the product development team had not had the opportunity to test the prototype with a vehicle. During the week of testing several software related problems were discovered, which required on-site software revisions. Consequently, the early runs in this program produced no viable data because of software bugs. The later data presented in this report is still prototype quality with incorrect software compensations applied. Results are preliminary to finishing the product development and should not be considered representative of the production model performance.



 $*$  ND-UV measures NO and NO<sub>2</sub> separately and combines these measurements to determine NO<sub>x</sub>

\*\* The OBS-1300 used for the BUG testing uses NDIR for hydrocarbons and a zirconia sensor for NO<sub>x</sub>;

The OBS-2200 used for the Chassis testing uses FID for THC and Chemiluminescence for  $NO<sub>x</sub>$ 

#### **Table 2-2. Description of Measurement Capabilities of Individual PEMS**

The fact that Table 3 shows diversity in PEMS product approach and offering is quite common at this stage of commercialization for an evolving product line. In fact, this diversity usually continues until standardization and specifications are established. For example, consider the CFR standard for measuring  $NO<sub>x</sub>$  in diesel exhaust, the measurement of oxygen in gasoline or any of the SAE protocols for vehicle interconnections. In these cases standardization took time and during that period the product performance specifications and operating tolerances were tightened as a result of the many discussions by the user and stakeholders groups. It is also common that during <span id="page-20-1"></span><span id="page-20-0"></span>standardization, that scientific approaches are narrowed and several manufacturers will offer products meeting standards but with their unique approach. Witness the number of companies offering  $NO<sub>x</sub>$  analyzers. Thus this research was important as it offered a quantitative review of current PEMS offerings and a forum for discussion of the future direction of performance standards.

## **2.4 First Comparative Test: Audit Bottle and Blended Audit Gas Checks**

The PEMS and FRM were compared over a series of concentrations delivered from an audit bottle. An additional blended gas check was also carried out during the stationary source/backup generator testing using audit bottle gases that were generated from a gas divider. The audit bottle and blended audit gas checks were done as blind comparisons with the concentration values not provided to the PEMS manufacturers until after the testing. As such, the analyzer ranges for the PEMS were not necessarily optimized for the concentration levels measured.

The audit bottle checks conducted during the BUG testing were with concentration levels representative of values typically found in the diluted exhaust as measured by UCR's FRM, rather than values in raw exhaust gas for which some PEMS were designed. Based on that observation, the audit bottles were designed before the mobile source tests so the audit bottle better represented the average concentrations that would be found in the dilute or raw exhaust depending on whether the PEMS was configured to sample a dilute or raw sample. All audit bottles were NIST traceable with  $\pm 1\%$  accuracy with a mixture of NO, THC, CO, and  $CO<sub>2</sub>$ .

During the BUG testing, measurements were also made at four concentration levels from a primary audit bottle diluted using a gas divider with  $\pm 1.5\%$  accuracy mass flow controllers. The four concentration levels were designed to span the ranges of concentrations that would be found in both raw and diluted exhaust.

## **2.5 Emission Measurement Tests –Backup Generator/Stationary Source**

In the first phase of testing, the flow rates and emission measurements from a backup generator were measured with four PEMS and compared with the values from the FRMs. The engine used for this testing was a model year 2000 CAT 3406C engine (serial number 4JK00740) that powered a backup generator rated at 350 kW. This same diesel engine was used in previous studies by UCR so there is a considerable record of the emissions history of the engine. Note the size of the engine was selected since it was similar in size to the CAT C-15 engine to be used in the second phase of the program. The engine was tested without a muffler. The experimental set-up for the BUG Testing is shown in Figure 2-1.



**Figure 2-1. Experimental Set-up for BUG Testing** 

The exhaust measurements were made on the BUG at four different load points at the rated speed, 1800 revolutions per minute (RPM). The load points were selected so as to represent a wide range of different flows and load conditions, both within and outside the NTE zones. This includes some test points standard and non-standard to the CFR test protocol, as UCR has published emissions at the standard load points. The actual test points are presented below:



During the testing, the engine was operated following the protocols specified in 40 CFR Part 89, except that the load points differed from the levels specified in the CFR. The engine was run at 100% load for approximately 30 minutes to precondition or warm-up the engine. Then the emissions were measured at the four-modes starting with the highest load point as per the protocol. The loading of the BUG was provided by a resistive load bank capable of providing 100% of the maximum load listed by the BUG manufacturer and power was calculated based on the amps and voltage measured for each leg of the 3 phase, 480 volt circuit. Data were collected at each mode for at least five minutes after the unit had stabilized.

To allow a determination of the precision of the measurement, tests were run seven times during each of the test periods. Flow and all exhaust emissions, including  $CO$ ,  $CO<sub>2</sub>$ , THC,  $NO<sub>x</sub>$ ,  $NO<sub>2</sub>$ , and PM mass, were measured with the CFR-specified laboratory instruments in UCR's MEL and with the PEMS instruments, depending on the capability of the individual PEMS. Each test run was set-up such that all of the PEMS units were sampling <span id="page-22-0"></span>simultaneously from the exhaust stream at the same time. The exhaust was sampled upstream of the inlet to the dilution tunnel used in the FRM. The volume sampled by the PEMS by the PEMS represented between 1% (at idle) to 0.25% (at full load) of the total exhaust volume and was not accounted for in the CVS calculations.

The BUG testing was conducted at two different times to accommodate as many PEMS as possible. PEMS 1, 2, and 3, were tested in November of 2004. PEMS 4 later became available and was tested in February of 2005. The measurements made with the FRM in February are slightly different from those of November, which can be traced to the load bank being changed from the November testing.

The fuel used for this testing was a CARB ultra low sulfur diesel (ULSD or sulfur <15 ppmw) fuel that was readily available through a local retail outlet.

## **2.6 Emission Measurement Tests –Heavy-duty Diesel Truck/Mobile Source**

Planning the second major phase of testing with the HHDDT/mobile source received much more attention than the first phase of testing with the Stationary engine as results were likely to aide in the ongoing discussions about a measurement allowance for the new EPA Changes to Test Procedures for Nonroad Engines and Heavy-duty Highway Engines. The new regulation covers measurement of emissions from in-use diesel engines, especially when the engine operated in the Not to Exceed (NTE) region of the engine map. The discussions soon pointed to the need for a developmental, low emission engine and Caterpillar offered a 475 hp Caterpillar C-15 ACERT engine with 200 hours or about 5,000 miles on it since being rebuilt. The engine was equipped with dual exhausts and a pair of oxidation catalysts and was certified to the 2.5 g/bhp-hr  $NO<sub>x</sub> + NMHC$  and 0.1 g/bhp-hr PM. The plan was to compare emissions measured with eight PEMS to values measured with the FRMs. Four of the PEMS measured only PM, three PEMS measured only gases and the remaining PEMS measured both gases and PM. However, one of the PM PEMS was not run within manufacturer's specifications and was excluded from the dataset. The chassis dynamometer tests were performed at CARB's heavy-duty vehicle laboratory (HDV lab) located in Los Angeles, CA.

The heart of the active testing phase was extensively discussed and those discussions resulted in a carefully planned series of steady state and transient driving schedules. These four driving schedules included:

- 1) Three-mode steady-state NTE test cycle run.
- 2) Stepped NTE test cycle
- 3) The Urban Dynamometer Driving Schedule (UDDS)
- 4) The 50-mph Cruise mode from the ARB 5-mode HHDDT test cycle

Test cycles 1) and 2) were specifically designed for the chassis dynamometer testing by members of the Calibrations and Standards task Force (CSTF) [now called the Emissions Measurement & Testing Committee], ARB, and UCR. The NTE driving schedules test cycles were designed to clearly delineate entry into, operation within, and exit, from the

<span id="page-23-0"></span>NTE-defined zone of engine operation. The steady-state NTE cycle was run at three different engine speeds with events in the NTE region. A second NTE cycle was designed with a "stepped" pattern, intended to simulate NTE-type vehicle/engine operation with gentle accelerations between modes and steady-state operation at each load point. More detail is provided below on the special NTE cycles.

The UDDS and 50-mph cruise driving schedules were intended to better represent 'realworld' operation, including stop-and-go transient operation and high speed cruise operation.

## **2.6.1 Steady-State NTE test cycle**

The steady-state NTE cycle was run with a vehicle speed of 40 miles per hour and included three different engine speeds and three different load points within each engine speed. The engine speeds were 1290, 1550, and 1770 rpm and the load points within each engine speed were 40%, 70%, and 100% load. Within each engine speed/load point, changes in grade were used to provide transitions into and out of the NTE zone. The determination of the specific engine speed and load points was done experimentally at the ARB HDV Lab in coordination with a Caterpillar representative. The changes in grade were repeated 7 times for each speed/load point within the cycle. Thus, each test run at a specific speed/load point includes 7 replicates of a specific NTE event. A description of the steady state NTE cycle at each speed is provided in Table 2-3. The cycle is shown graphically in Figure 2-2 using load data broadcast from the ECM via J1939 protocol. The NTE steady state cycle was repeated twice for each engine speed, with one run each on two different days.



**NTE Steady State Cycle - 1770 RPM Run #1** 



<span id="page-25-0"></span>





**Table 2-3. Description of Steady state Cycle** 

#### **2.6.2 Stepped NTE cycle**

<span id="page-26-0"></span>The stepped NTE test cycle consists of seven separate modes at four different vehicle speeds that "stair-step" increases in vehicle speed from 10 mph to 60 mph, and then decreases in vehicle speed back to 10 mph in the reverse order. Each mode is separated by an acceleration or deceleration event. This pattern is repeated twice within each cycle. The test schedule is provided in Table 2-4, including vehicle speed, grade, engine speed, and load. The grade was fixed for this cycle to maintain the appropriate load. The cycle is shown graphically in Figure 2-3 using loads based on J1939 data. The intent of the cycle is to exercise the vehicle/engine in a predictable and repeatable manner. The determination of the specific engine speed and load points was done experimentally at the ARB HDV Lab in coordination with a Caterpillar representative. Two sets of three repetitions of the entire stepped NTE cycle were performed on two different days.



**Table 2-4. Description of Stepped NTE Cycle.** 

#### **NTE Stepped Cycle Run #1**

<span id="page-27-0"></span>

**Figure 2-3. NTE Stepped Cycle Plot** 

#### **2.6.3 Emissions inventory test cycles**

Two existing dynamometer driving schedules were utilized to try to better represent realworld engine/vehicle operation. The first test cycle is the federal heavy-duty vehicle Urban Dynamometer Driving Schedule (UDDS) and is often used to collect emissions data on engines already in heavy, heavy-duty diesel (HHD) trucks. The test cycle is a transient test cycle with a short cruise section, and hence exercises both the test vehicle and PEMS over a fairly wide range of operation. This cycle covers a distance of 5.55 miles with an average speed of 18.8 mph and maximum speed of 58 mph. This cycle is shown in Figure 2-4.

The second test cycle is a "real-world", emissions inventory cycle, the ARB-developed "50 mph Cruise Mode" which is part of the ARB 5-mode heavy-heavy-duty diesel vehicle test cycle. In addition to including some operation within the NTE-zone, this test cycle exercises the PEMS in a manner similar to what would actually occur if the vehicle were operated over-the-road under highway cruise conditions. The test cycle has a cycleaverage-speed of 50.4 mph, a maximum speed of 67.1 mph, is about 12.6 minutes in duration, and is about 10.6 miles long. This cycle is shown in Figure 2-5.



<span id="page-28-0"></span>



**ARB 50 mph HHDDT Cruise mode** 



**Figure 2-5. 50 mph Cruise Cycle** 

#### **2.6.4 Test weight and road-load horsepower**

For the emissions inventory test cycles (UDDS and 50-mph Cruise mode) the test weight was set at approximately 53,000 lbs. This test weight is similar to the gross vehicle weight of the CE-CERT mobile emissions lab plus the weight of the test vehicle. This provided test conditions similar to those that will be utilized during future on-road testing. This value is also similar to the average load for HHD trucks on California roads. For the three-mode steady-state test cycle and the four-speed, seven-mode steady state test cycle, the dynamometer settings were determined iteratively, based on achieving the desired engine speeds, loads and vehicle speeds. The experimental set-up for the chassis de a contract of the contract of the dynamometer testing is shown in Figure 2-6.<br>
de a contract of the emissions inventory test cycles (UDE)<br>
was set at approximately 53,000 lbs. This<br>
weight of the CE-CERT mobile emission

<span id="page-29-1"></span><span id="page-29-0"></span>

**Figure 2-6. Experimental Set-up for the Chassis Dynamometer Testing** 

## **2.6.5 Engine/vehicle preconditioning**

The vehicle was "soaked" (parked) in the chassis dynamometer test cell, which was not temperature controlled, prior to testing and between each of the test days. The test vehicle was fully warmed up to its normal operating temperature at the start of each test day using the test cycle to be used that day, followed by steady-state operation at speed of 50 mph for a period of ten minutes. The first test cycle for the day was run immediately after this sequence.

In order to permit PM filter changes, monitor PEMS operation, etc., a period of 30 minutes was included between each run. In order to improve repeatability, the 10 minutes- at-50-mph preconditioning was performed between each successive run.

## **3.0 Emissions Results – Stationary Source/Backup Generator**

<span id="page-30-1"></span><span id="page-30-0"></span>The PEMS tested for the BUG were PEMS1 (Montana System), PEMS2 (Semtech D system), PEMS3 (the RAVEM system), and PEMS4 (the OBS-1300).

#### **3.1 Audit Bottle and Blended Gas Checks**

A primary audit calibration bottle was used as a checkpoint between the PEMS and the FRM. The bottle was NIST traceable with  $\pm 1\%$  accuracy with a mixture of NO, THC,  $CO$ , and  $CO<sub>2</sub>$ . The percent differences between the PEMS and the FRM are shown below in the Figure 3-1 and details provided in Table 3-1. As noted earlier, the concentration ranges in the audit gas were representative of those found in dilute exhaust measurements and not suitable for some of the PEMS measuring raw exhaust concentrations. For example, PEMS2 manufacturer indicated the values were below the design specifications for their instrument. PEMS3 did not measure THC and no THC data were available from PEMS2. The audit bottle concentrations for PEMS4 also showed considerable deviation for NO. Note that the audit bottle tests with PEMS4 were completed after the testing on the BUG and the manufacturer of PEMS4 noted that after the exhaust test measurements, but before the audit, the zirconia sensor for  $NO<sub>x</sub>$  was exposed to water. They suspected the water exposure damaged the unit, leading to potentially erroneous readings of the audit bottle and the calibration gas checks below. Since the unit was damaged after the BUG testing, only the audit bottle measurements would have been affected. It should be noted that hydrocarbons for this section are denoted as THC, although two PEMS utilized NDIR for these measurements.



**Figure 3-1 Percent Differences from the Primary Audit Bottle. Audit bottle concentrations included in the Figure.** 

<span id="page-31-0"></span>Except for the previously noted differences, all instruments measured  $CO<sub>2</sub>$  and NO levels in primary audit gas within  $\sim$  5%. The best agreement is found with the FRM instruments, including an exact match with the  $CO<sub>2</sub>$  value.

It is important to recognize that the concentrations of the primary audit gas were more similar to the values found in diluted streams rather than raw. Since most of the PEMS surveyed are designed to measure the emissions in raw gas streams, differences in concentrations at these levels would not necessarily indicate that a similar difference would be found at the higher concentration levels found in raw exhaust. The audit bottle concentration measurements are compared with the range of concentration values observed in the raw and dilute exhaust in Table 3-1.

The differences for CO were not included in Figure 3-1 as some of the values would skew the presentation of the data in the figure. For CO, the differences were -40% for PEMS1, 59% for PEMS2 and 1.4% for the FRM. PEMS3 and PEMS4 did not provide CO data.

			<b>Concentration Measured</b>				
ID	Cal Type	Date	NO ppm	THC ppmC1	CO ppm	CO <sub>2</sub> %	
PEMS <sub>1</sub>	<b>Bottle</b>	2-Nov	149.5	31.00	15.00	1.440	
PEMS <sub>2</sub>	<b>Bottle</b>	2-Nov	161.0	-	40.00	1.640	
PEMS <sub>3</sub>	<b>Bottle</b>	2-Nov	160.0			1.533	
<b>FRM</b>	<b>Bottle</b>	4-Nov	150.5	24.67	25.49	1.556	
PEMS <sub>4</sub>	<b>Bottle</b>	3-Feb	249.0	33.00	۰	1.600	
<b>FRM</b>	<b>Bottle</b>	3-Feb	149.2	23.82	25.13	1.573	
Audit			151.8	23.91	25.14	1.556	
<b>Raw Exhaust</b>		$250 - 1740$	$20 - 101$	70 - 251	$2.8 - 8.4$		
Dilute Exhaust			$53 - 297$	$6.0 - 18$	$29 - 68$	$0.5 - 2.6$	

**Table 3-1 Comparison of PEMS & FRM Measurement of 1% NIST Audit Bottle** 

## **3.2 Blended Audit Gas Check**

Another set of tests were conducted using audit gases prepared using a gas divider with ±1.5% accuracy mass flow controllers (MFCs) and included a zero and four span ranges. The span gases were measured in the order of high to low concentration, with the higher concentrations intended to be closer to those expected for raw exhaust measurements (span 3 and 4), and the lower concentrations being more representative of the dilute concentrations measured with the FRM. Figure 3-2 below shows the percent difference between the measurements and the  $CO<sub>2</sub>$  and NO calibration gases for the measurements. Note PEMS4 was tested in February and slightly higher calibration gas levels were used for  $CO<sub>2</sub>$  and NO at that time. The concentrations provided on the right are from the February tests where there are differences from the values used in November. A duplicate run was conducted for the FRM during the November test period.



**Figure 3-2 Percent Difference Between Measured Value & Calibration Gas Concentration a) NO and b) CO2. Spans 3 and 4 were Close to the Raw Values Expected for a BUG.** 

As is evident in Figure 3-2, the percent difference between the true value and that measured by the FRM ranged from zero to a maximum of  $1.8\%$  for both NO and CO<sub>2</sub> gases. For the PEMS, the difference depended on the concentration of the blended audit gas. Further, some data in Figure 3-2 has all of the deviations to one side of the true value and is suggestive of bias for some PEMS. The question of bias will be addressed later.

An important point for proper analysis of the results plotted in Figure 3-2 is the relationship between the concentration of the audit gas and the design specification for the analyzer. For example, the differences for both gases for Spans 3 and 4 are generally within  $\pm 10\%$ . For Spans 1 and 2, where the concentrations generally are outside the range of specification for the raw gas analyzers, the deviation ranged from  $-15\%$  to  $+30\%$  for  $CO<sub>2</sub>$  and from -15% to +27% for NO. This finding is not surprising as error generally increases when an instrument is operating near its lower detection limit. The point about the audit gases being outside the design range of the analyzers is well represented in a set of charts from one of the PEMS manufacturers. Note in Figure 3-3 that all the blended audit gases for the  $CO<sub>2</sub>$  are below the measured values during the testing while for NO only two of the blended audit values fell within the test range.

<span id="page-33-1"></span><span id="page-33-0"></span>

Figure 3-3 Comparison of Blended Audit Gases & Engine Emissions for CO<sub>2</sub> and NO<sub>x</sub>

The results for CO and THC are not shown in the graph but values can be viewed in the detailed results of Appendix C. Some of the percent differences from the true values are rather high and most of these values suffer the problem of having the concentration outside the design value for the PEMS.

## **3.3 Exhaust Flow Rates**

Both concentrations and flow rates must be accurately measured in order to obtain accurate emission rates. Exhaust flow rates for the FRM and PEMS 1, 2 and 4 were compared. The PEMS3's design uses a proportional partial flow dilution system, instead of direct measurement of the flow rate. Emission rates are determined using the ratio of the sampling probe diameter to the exhaust diameter.

Figure 3-4 below compares flow rates provided by the manufacturers of PEMS1, 2 and 4 with the flow rates measured with the FRM over the range of 420 to 1000 standard cubic feet per minute (SCFM) for the four load points used in the study. The PEMS flow rates were obtained from calculation (PEMS1) or direct measurement of exhaust flow (PEMS2 and 4). A description of the flow measurement method used in UCR's MEL is provided in Appendix A. MEL's primary dilution system is configured as a full-flow constant volume sampling (CVS) system with a smooth approach orifice (SAO) venturi and dynamic flow controller. The SAO venturi has the advantage of no moving parts and repeatable accuracy at high throughput with low-pressure drop. Tunnel flow rate is adjustable from 1000 to 4000 scfm with accuracy of 0.5% of full scale. The FRM exhaust flow was determined by subtracting the measured flow of the dilution air from the total CVS tunnel flow. A paired t-test showed that all flow differences between the FRM and the PEMS were statistically significant except for PEMS1 at the highest flow rate.

<span id="page-34-0"></span>

**Figure 3-4 Relative Difference in Flow Rate Between FRM and PEMS** 

An experimental note is that the thermocouple for exhaust temperature measurement for PEMS2's flow measurement device was limited to 400°C, so the as supplied unit could not accurately compute the exhaust density and standard volumetric flow when the exhaust was above this temperature. UCR worked with the PEMS2 manufacturer to provide exhaust temperature during the BUG testing. PEMS2 subsequently used this supplied temperature and actual volumetric flow from their flow device during the entire test to compute standard volumetric flow and to calculate their mass emissions.

Another perspective of the quality of the data over the flow range is the trend analysis shown in Figure 3-5. From this analysis, the three measuring devices are very linear over the measured range as evident from the high values (closeness to 1) for the coefficient of determination ( $\mathbb{R}^2$ ). One consequence of the close agreement for flow of PEMS 1, 2 & 4 and the FRM is that any subsequent differences observed in measured emissions rates are likely due to concentration issues rather than flow measurements.

<span id="page-35-1"></span><span id="page-35-0"></span>

**Figure 3-5 Correlation Plot for PEMS and FRM for Exhaust Flow Rate** 

#### **3.4 Mass Emission Rates**

BUG emission rates for each gas were calculated in grams per hour  $(g/hr)$  and the percentage difference relative to the FRM value is shown in the figures below. It was decided to evaluate the percentage differences and correlations on a g/hr basis since this is the purer measurement of the ability for the PEMS to make mass measurements over a specified cycle or time period than grams per horsepower-hour (g/hp-hr). Specifically, on a g/hp-hr basis, the highest emission rates are found at the load points where the horsepower is very low, resulting in an overemphasis of these values when looking at correlations between the PEMS and FRM. Measurements on a g/hp-hr basis are discussed below since the g/hp-hr measurements are important in understanding the measurement levels in relationship to the certification standards and applicability for NTE measurements. The g/hp-hr values for the FRM are also included in the figures relating the g/hr percentage differences. Note that results for the mass emissions in g/hr or g/hp-hr depend on the independent measurements of mass flow and concentration so errors in flow or concentration will carry through into the calculated mass emission rate. An alternative method for examining the data is to normalize the emissions rates for the regulated emissions to  $CO<sub>2</sub>$ , which would automatically compensate for any differences in flow rate, as shown below. Detailed BUG emissions rates for the FRM and each PEMS are provided in Appendix D in both g/hr and g/hp-hr units. Statistical comparisons of the BUG emission rates in g/hr units are provided in Appendix E.
Figures 3-6a and 3-6b show the results for  $CO_2$  and  $NO_x$  when the PEMS are compared with the FRM. The agreement for was  $CO<sub>2</sub>$  relatively good for PEMS1, 2, and 4, with PEMS4 having the highest overall difference, about 10%. The agreement of PEMS3 was good at the highest flow rates, but the difference was about 50% at the lowest load. This difference was identified as being related to a failed component used in controlling flow. Similar deviations are also found for PEMS3 for the other emissions at the lower load points. It is interesting to note that the FRM measured higher  $CO<sub>2</sub>$  emissions than all PEMS. For  $CO<sub>2</sub>$  emission rates, the differences between the FRM and the PEMS were statistically significant for all comparisons except PEMS 1, 2, and 3 at the highest load point and PEMS4 at the lowest load point.



Figure 3-6 Mass Emissions (gm/hr) for PEMS Relative to FRM a) NO<sub>x</sub> and b) CO<sub>2</sub> Emission **Rates in gm/hp-hr provided for FRM.** 

 $NO<sub>x</sub>$  values showed agreement within ~10% or better for PEMS2 and 4 with the FRM. PEMS3 values showed good agreement at high loads, with larger errors at the lowest load. For PEMS3, the  $NO_x/CO_2$  ratios, which eliminate the effects of flow measurement, were within 10% of those measured by the FRM for all load points. PEMS1 values were 12% to 30% higher than the FRM values. Some of the differences in  $NO<sub>x</sub>$  emissions for the PEMS1 are related to the omitted humidity correction  $(\sim 10\%)$  and a bias of about the same magnitude observed with the calibration gases, since the flow rates show good correlation with the FRM.

Figure 3-7 shows the difference in emission rates for THC and PM. Three manufacturers - PEMS1, PEMS2, and PEMS4 - reported hydrocarbons (HC) and resultant values were significantly different relative to the FRM. It should be noted that the actual HC concentrations are relatively low, even in comparisons with the upcoming proposed NTE standards (as discussed below). Since PEMS1, 2 and 4 had accurate flow measurements, the source of the difference must be in the measured concentration. The manufacturer of PEMS2 said the values in the raw gas were outside the range the instrument was set up to measure. The HC measurements for PEMS1 and 4 were measured with NDIR.

Two manufacturers, PEMS1 and PEMS3, measured particulate matter (PM). For PEMS1 results were significantly lower than the FRM at all ranges. For PEMS3 values were within  $\sim$ 20% of the FRM at higher loads, with larger deviation at the lower loads. PM/CO<sub>2</sub> ratios for PEMS3 for the 5 and 25% load points were 22-25% lower than those measured by the FRM.



**Figure 3-7. PEMS Mass Emissions Relative to FRM at Several Loads (a) THC and (b) PM. Emission Rates in gms/hp-hr provided for FRM.** 

Plots showing the correlation over a broad range are provided below in Figures 3-8 (a) (d).



Figure 3-8 Plot of Correlation between FRM and PEMS for (a) NO<sub>x</sub> (b) CO2 (c) PM & (d) THC

#### **3.5 Results in Grams per Horsepower Hour**

Figure 3-9 was developed to show the comparative emission rates on a g/hp-hr basis for  $NO<sub>x</sub>$ , THC, and PM for the FRM measurements and each PEMS. The error bars on the graphs represent 95% confidence intervals. The emission rates were calculated using the measured mass emission rates and electrical power output from the load bank. The emissions rates and associated error bars for the 5% load point for THC and PM are divided by 5 to allow presentation of all modes on the same scale. One advantage of Figure 3-9 is that it visually allows a perspective on the closeness of the measurements among the different methods and the significance of the 95% confidence limits. Another feature in Figure 3-9 shows the emissions rate per unit of work increasing rapidly with at the lowest power as is usually reported.



Figure 3-9 Comparison of g/hp-hr Emission Rates and 95% Confidence Limits. (a) NO<sub>x</sub>, (b) **THC, (c) PM** 

## **3.6 Data Analysis within the NTE Zone**

As explained earlier, the original intent of the project was to broadly survey the accuracy of the PEMS when applied to data gathering for either NTE or inventory purposes. However, as the project developed, a greater interest was assigned to defining the accuracy and percent deviation of the PEMS within the NTE zones. The NTE zones were defined in earlier sections.

Even though the engine tested in the backup generator did not need to comply with NTE regulations, we thought it was of value to examine the points operating within the NTE zone, namely at 65% and 100% power. To gain some insight as to the closeness of the PEMS to the FRM within the NTE zone,  $EPA<sup>3</sup>$  provided values for the NTE limits so comparative analyses could be made of PEMS and the FRM. Table 3-2 below compares both the absolute difference of the emissions rates determined by the FRM and by the PEMS and the relative error when the absolute difference is divided by the NTE threshold limit. Because it might be close on a brake-basis, we included the point at 25% power as well.



Differences represent the absolute value of the difference between the FRM and PEMS

Considering only the 65% and 100% load points in this analysis, the absolute differences between the FRM and the PEMS  $NO<sub>x</sub>$  measurements were 4 to 23% of the NTE limit for PEMS2, 3 & 4, with significantly larger differences found for PEMS1. For THC, PEMS2 was most comparable with the FRM with the differences relative to the NTE limit being 11 and 12% while PEMS1 and 4 had relative differences ranging from 41 to 75%. For PM, PEMS1 and 3 showed differences of between 65 to 79% and 214 and 261%, respectively, from the FRM in comparison with the 0.03 g/hp-hr NTE standard that might be applicable in the future.

The goal in this section on the NTE perspective was to present an analysis that might provide some insight as to the allowable tolerance that is needed when enforcement of the NTE rule is based on in-use measurements with PEMS rather than with laboratory

**Table 3-2 Comparison between Percentage Differences on g/hp-hr Basis Relative to NTE**  Limits Provided by the EPA<sup>2</sup>

measurements and instruments. The analysis suggests that tolerances will depend on the PEMS unit and on what pollutant is being measured. For PM measurements, it appeared from these analyses that additional developments will be required to provide techniques suitable for PEMS/NTE testing.

### **3.7 Results Based on Fuel Consumed or CO2**

Emission rates per unit of fuel consumed or on a  $CO<sub>2</sub>$  basis are of interest for those involved with either/both determining emissions inventories or enforcing the not-toexceed regulations. Harley and co-workers<sup>4</sup> shows that one advantage is a quick determination of the total emissions if the fuel consumed in a specific area is known. Another advantage of this approach is that calculations based on ratios of emissions to  $CO<sub>2</sub>$  are independent of exhaust flow rate. Thus determining the ratio of emission rates relative to the  $CO<sub>2</sub>$  rates can be useful in identifying whether the source of the difference between the FRM and PEMS is related to the measurement of concentration or flow rate. Figure 3-10 was provided to illustrate when the rates of  $NO<sub>x</sub>$ , THC, and PM is divided by the rates of  $CO<sub>2</sub>$ .

This approach and deeper analysis was particularly useful for the data from PEMS3, especially at the two lowest load points where the manufacturer indicated that the partialflow sampling system appeared to be under-sampling by approximately 55% at the lowest load point and by approximately 12% at the second lowest load point. The manufacturer suspected that a differential pressure sensor, used to maintain isokinetic sampling, had malfunctioned during the testing and this malfunction lead to the undersampled flow rates. Note in Figure 3-10 that the differences between the FRM and PEMS3 were considerably narrowed, confirming that the flow measurement error had propagated into the calculation of the emission rate. For example, the difference between the FRM and PEMS3 for the measurement of  $CO<sub>2</sub>$  and  $NO<sub>x</sub>$  at the lowest power was reduced from 50% to about 10%.

Several other artifacts were noticed. For example, the  $NO<sub>x</sub>/CO<sub>2</sub>$  emissions for PEMS1 were biased approximately 5 to 10 % higher relative to the FRM. This difference could be related to the concentration measurements and should be further investigated. For PEMS2, a slightly higher bias in  $NO_x/CO_2$  ratios relative to the FRM was found compared to the straight  $NO<sub>x</sub>$  emissions. This is probably due to the lower  $CO<sub>2</sub>$  emissions in comparison with the FRM at the three lowest load points. Similarly, for PEMS4, the biases in the  $NO_x/CO_2$  ratios were higher (65% and 100%) or less negative (25%) compared to the FRM than the straight  $NO<sub>x</sub>$  emissions at the three highest load points, due to lower  $CO<sub>2</sub>$  measurements compared to the FRM at those points.



Figure 3-10 PEMS Emission Rate Relative to CO<sub>2</sub> When Compared with FRM Emission Rate Relative to CO<sub>2</sub>. For: a) NO<sub>x</sub>, b) PM, c) THC.

### **3.8 Precision**

A definition of precision can be found in ASTM E-177: *Standard Practice for the Use of the terms Precision and Bias in ASTM Test Methods.* Therein precision is defined as a concept related to the closeness of agreement between test results obtained under like conditions when measurements are made for a test method in a state of statistical control. For this report, the precision of the different PEMS and the FRM were calculated using the coefficient of variation (COV) at one standard deviation derived from the average value of the seven repeat iterations at each load point. Plots of the results are shown in Figure 3-11 below.

Of the emission components,  $NO<sub>x</sub>$  and  $CO<sub>2</sub>$  showed the lowest COVs or scatter across all of the measurement techniques. A one-tailed F-test showed no statistically significant differences in the precision of the  $NO<sub>x</sub>$  measurements between the FRM and PEMS 2 and 4. The COVs for  $NO<sub>x</sub>$  for the FRM were lower at a statistically significant level relative to PEMS 1 and 3, for all but the lowest load point for PEMS 1. For  $CO<sub>2</sub>$ , the COVs for most PEMS and the FRM were within 2% for a majority of the test points. The COVs for the FRM were slightly lower than those for the PEMS, with this result being statistically significant for all but the 5 ( $p=0.056$ ), 25 and 100% load points for PEMS 4.

For PM, the COV or precision of the measurements was generally below 10% for both the FRM and PEMS, with no consistent advantages in precision seen for the FRM. However, recall the disparity in the accuracy for the other PM methods over all ranges. THC measurements generally showed higher dispersion or COVs for the PEMS than the FRM. The COV for THC with the FRM ranged between 4 and 7 %, while the COV for PEMS1 was from 17-33%, the COV for PEMS2 was from 7-20%, and the COV for PEMS4 was from 23-37%. The COV for CO emissions measurements was generally below 10% for the FRM, PEMS1 and PEMS2. No statistically significant differences in the COVs were found for PEMS1 or 2 for CO, except for PEMS2 at the lowest load point. The CO COVs for PEMS 4 and PEMS3 at the low load points were larger than those for the FRM or the other PEMS.

The significant observation from the precision analyses is that precision and accuracy are independent measures. In fact, the data indicate that measurements for PM and THC were quite inaccurate but highly precise.



Figure 3-11 Precision of the FRM and PEMS for the Various Measurements. a) CO<sub>2</sub>, b) NO<sub>x</sub>, **c) PM, and d) THC** 

# **4.0 Emissions Results – Mobile Source/Heavy-duty Truck Testing - Gas-Phase Emissions**

The PEMS tested for on the chassis dynamometer for gas phase emissions were PEMS1 (Montana System), PEMS2 (Semtech D system), PEMS3 (the RAVEM system), and PEMS4 (the OBS-2200). In response to strong urging from both EPA and ARB management, Horiba submitted a prototype OBS-2200 instrument with unproven software for the chassis dynamometer test program at the ARB laboratory in Los Angeles. The product development team for this new instrument had not had the opportunity to test the prototype with a vehicle before shipping it to the ARB laboratory. During the week of testing several software related problems were discovered, which required on-site software revisions. The early runs in this program produced no viable data because of software bugs. The later data presented in this report is still prototype quality with incorrect software compensations applied. These results are preliminary to finishing the product cycle and should not be considered representative of the production model performance.

### **4.1 Primary Audit Bottle Check**

The percent differences from the audit bottle for the FRM and PEMS during the chassis dynamometer testing are shown below in Figure 4-1, along with the raw and dilute audit bottle concentrations. Each PEMS was tested with either the raw or dilute audit bottle depending on whether the PEMS was configured to sample dilute or raw exhaust. The detailed results are provided in Appendix C. PEMS3 did not measure THC. It should be noted that the audit bottle concentrations were not provided to the PEMS manufacturers prior to the audit bottle check or before the engine testing, hence, the analyzer ranges were not necessarily optimized for the given concentration values during the engine testing.

The deviation from audit bottle concentration varied by PEMS and by pollutant. For the FRM all concentrations were within 2.5% of the audit bottle concentrations for both raw and dilute concentration levels. For PEMS2, the measured concentrations were within 2% of the audit bottle for  $NO<sub>x</sub>$  and  $CO<sub>2</sub>$ , with larger differences for THC (-7%) and CO  $(+13%)$ . PEMS3 measured the same concentration as the audit bottle for NO<sub>x</sub>, with differences of  $-11\%$  and  $-8\%$ , respectively, for CO and CO<sub>2</sub>. PEMS4 showed the largest deviations from the audit bottle concentrations with differences of 26% and 17% for NO and  $CO<sub>2</sub>$ , respectively, and differences of 20% and  $-9\%$ , respectively, for THC and CO. The manufacturer of PEMS4 indicated that software bugs in their prototype unit caused reading and logging errors that contributed to the deviations from the audit bottle on Tuesday during this test period. The software was updated prior to conducting the emissions tests presented for PEMS4 later in the section. It should be noted that hydrocarbons for this section are denoted as THC, although one PEMS utilized NDIR for these measurements.



**Figure 4-1 Percent Difference between PEMS/FRM and the Audit Bottle.** 

### **4.2 Integrated Cycle Emissions and Exhaust Flow Rates**

Mass emission rates in grams per cycle and exhaust flow rates were determined for each of the cycles integrated over the entire duration of the cycle. This included the UDDS and 50 mph cruise cycles, the steady-state NTE cycle and the NTE stepped cycles. For the steady state and stepped NTE cycles, the integrated results were collected over all NTE events within a particular cycle and the transitions between NTE events. For the NTE steady-state cycles, this included all 7 values measured at the NTE events at the 40%, 70%, and 100% loads for one test iteration. The total number of complete cycles conducted during the testing program was as follows: seven UDDS cycles, seven 50 mph cruise cycles, six NTE stepped cycles and two each of the full NTE 1290, NTE 1500, and NTE 1770 cycles. Data for PEMS1 and PEMS4 are only available for one test for each of the NTE 1290, NTE 1500, NTE 1770. Only one test is also available for PEMS1 on the NTE stepped cycle and for PEMS4 on the 50 mph cruise cycle. There are no error bars in the Figures for PEMS/test cycle combinations where only a single test is available.

### **4.3 Integrated Exhaust Flow Rates**

Both concentrations and flow rates must be accurately measured in order to obtain accurate emission rates. The average exhaust flow rates were determined for each of the integrated cycles. The FRM exhaust flow was determined by subtracting the measured flow of the dilution air from the total CVS tunnel flow, as explained in an earlier section.

The PEMS flow rates were obtained from their respective direct measurement of exhaust flow, except for the PEMS3 approach. PEMS3 does not require the direct measurement of flow rate as long as their partial dilution system maintains isokinetic, and therefore

proportional, sampling of the main exhaust stream. For PEMS3 emission rates are determined using the ratio of the sampling probe diameter to the exhaust diameter.

Figure 4-2(a) below is a comparative plot of the flow rates measured by the FRM and PEMS 1, 2 and 4. The exhaust flow rates measured by the FRM ranged from 350 to 800 standard cubic feet per minute (SCFM). Exhaust flow rates for PEMS1 were approximately 10% different from those of the FRM. Exhaust flow rates for PEMS2 ranged from 0 to 4% with an average deviation of 2.7% as compared to those measured by the FRM. PEMS4 also showed good agreement with the FRM, with a range from 0 to 4% and an average deviation of less than 1%.

A different perspective of the correlation between the whole range of the exhaust flow rates for the FRM, PEMS2, and PEMS4 is provided in Figure 4-2(b). The curve fit or coefficient of determination over the whole range is excellent (>99%), although there are differences in the slopes and intercepts.



**Figure 4-2 (a) Comparative FRM and PEMS Integrated Exhaust Flow Rates for and (b) Correlation Plot** 

### **4.4 Integrated Mass Emission Rates**

Detailed mass emission rate results are provided in Appendix F, along with the average concentration measurements for the tests for each instrument. The statistical analysis results for the integrated comparisons are provided in Appendix G.

The integrated results for  $NO_x$  and  $CO_2$ , from the FRM and PEMS are presented in Figures 4-3 and 4-4. Error bars shown represent the 95% confidence limits based on results from multiple runs. For the NTE steady state and NTE stepped cycles, the integrated values represent the emissions collected over the entire duration of the cycle, including all speed and load points within the cycles (see earlier Figures 2-2 and 2-3). Except for PEMS1, the  $NO<sub>x</sub>$  readings for the PEMS were higher than those obtained by the FRM. PEMS2 readings ranged from 5 to 12% higher than the FRM over the different cycles. For PEMS3,  $NO<sub>x</sub>$  emissions ranged from 14 to 21% higher than the FRM. PEMS4 showed differences in  $NO<sub>x</sub>$  emissions generally ranging from 19 to 40% higher than the FRM. It should be noted that some of these differences might be related to the calibration

differences seen in the primary audit bottle concentrations and associated software problems. The manufacturer of PEMS4 indicated software issues in their prototype unit contributed to the observed differences. For PEMS1,  $NO<sub>x</sub>$  values ranged from 13% lower to 25% higher than the FRM. A statistical analysis indicates that the differences between the FRM and PEMS for  $NO<sub>x</sub>$  are statistically significant at a 95% confidence level for most cycles for the most PEMS. Exceptions include the results for the NTE 1290 which are not statistically different, the results for the NTE 1770 cycle that are statistically significant at only a 90% confidence level, and the results for PEMS1 on the UDDS and 50 mph cruise cycles.



Figure 4-3. Mass Emissions (NO<sub>x</sub> g/cycle) for PEMS Relative to FRM

 $CO<sub>2</sub>$  emissions also showed a trend with the PEMS having generally higher emissions than the FRM. The PEMS2  $CO<sub>2</sub>$  emissions were within 5% of those measured by the FRM for most of the test runs, with no statistically significant differences for NTE 1290, 1500 and 1770 steady state cycles. These differences are comparable to those found between the exhaust flow rates for the FRM and PEMS2, as discussed below. The differences in  $CO<sub>2</sub>$  for PEMS3 were slightly higher ranging from 9 to 14% over the difference cycles. The  $CO<sub>2</sub>$  emissions for PEMS4 were between 11 and 26% higher than the FRM. For comparison, the calibration for PEMS4 was approximately 17% high on the audit bottle. The  $CO<sub>2</sub>$  emissions measurements for PEMS1 were below those of the FRM for all test cycles, with a range from 4 to 35% lower.



Figure 4-4 Mass Emissions (CO<sub>2</sub> g/cycle) for PEMS Relative to FRM.

The percentage differences for the integrated cycles are provided in Table 4-1 for all emissions.

	THC (g/cycle)			CO (g/cycle)			$NOx$ (g/cycle)				$CO2$ (g/cycle)			
						PEMS								
			4							4				
<b>UDDS</b>	-103	57	76	-42	20	$-29$	$-13$	12	21	25	-35		9	13
50 mph cruise	-78	-3	-6	$-16$		$-25$	-9		16	38	$-19$	21	14	22
<b>NTE 1290</b>	-88	-36	-4	$-10$	51	-38	5	9	16	40	$-16$	5.3	13	21
<b>NTE 1550</b>	-46	-33	$-29$	$-16$	36	$-29$	13		15	37	-14	2.2	13	26
<b>NTE 1770</b>	-83		-61	-8	38	$-25$	25	5	14	30	-4	21	12	11
<b>NTE Stepped</b>	-4	-18		-25	43	-64	-8	8	18	19	-14		13	19

**Table 4-1. Percentage Differences for Integrated Cycle Results (in percent)** 

To better understand the differences in the independent factors of concentration and exhaust flow, the mass emission rate of  $NO<sub>x</sub>$  was divided by that of the  $CO<sub>2</sub>$  to provide a metric that is nearly independent of flow rate. The results are provided in Figure 4-5. The  $NO<sub>x</sub>$  emissions normalized to  $CO<sub>2</sub>$  for PEMS2 show slightly closer agreement with the FRM, now ranging from 3 to 10%. For PEMS3, the agreement with the FRM improved considerably and was within 5% for most cycles, except the UDDS which was 10%. This indicates that differences in flow measurements could be a primary source of the differences between the FRM and PEMS3. The agreement between PEMS4 and the FRM also improved and now ranged from 0 to 17%. While part of the narrowing in deviation could be attributed to flow, the exhaust flow measurements for PEMS4 are similar to those of the FRM. It is more likely the fairly large differences found for PEMS4 for  $NO<sub>x</sub>$ and  $CO_2$  were simply normalized using this approach. For PEMS1, the  $NO_x/CO_2$  values were all higher than those of the FRM, probably due in part to the lower  $CO<sub>2</sub>$  emissions reported by PEMS1.



Figure 4-5 NO<sub>x</sub>/CO<sub>2</sub> Emission Rates for the FRM and PEMS.

The integrated THC and CO measurements for the FRM and PEMS are shown in Figure 4-6a and 4-6b, respectively. PEMS3 did not measure THC and CO emissions were not available for PEMS4 due to software issues. As such, they are not included in those respective figures.

The measurement of THC emissions showed greater variability than measurements of other emissions components. Under the test conditions and with the equipment used for this program, the percentage differences between the FRM and PEMS were generally not statistically significant. Interestingly, the approximately 20% higher bias in the THC calibration gas for PEMS4 is not seen in the mass emission results which are lower than the FRM values for all but the UDDS and 50 mph cruise cycles. THC levels for PEMS1 were well below those of the FRM. The THC levels measured in this study were relatively low overall  $(\sim 10 \text{ ppm}$  and were below the NTE standard thresholds. Hence, the percentage differences relative to the NTE standards were generally less than 5%, as discussed below.

The CO emissions for PEMS2 and PEMS3 show opposite trends in comparison with the FRM. CO emissions for PEMS1 were generally lower than those for the FRM, ranging from 8 to 42% lower. PEMS2 CO emissions were higher compared with the FRM for all test cycles. CO emissions for PEMS2 ranged from 20 to 50% higher than the FRM. Some of the difference can be attributed to the  $\sim$ 13% higher readings on audit gas. The measured CO emissions for PEMS3 were all lower than those for the FRM with the differences ranging from 25 to 64%. Some of this difference could be attributed to the approximately 11% low audit bottle reading that was found for PEMS3 on CO. To keep the CO measurements in perspective, the levels measured in this research were ~100ppmv and well below the NTE standard. Hence, the percentage difference relative to the NTE standards were generally less than 2% as discussed below.



**Figure 4-6 FRM and PEMS Mass Emissions Rates for Various Cycles (a) THC and (b) CO.** 



Correlation plots for  $CO_2$ ,  $NO_x$ , PM, and THC between the PEMS and FRM were also developed for the integrated data. These plots are shown in Figures 4-7 (a) to (d).

Figure 4-7 Correlation Plots of FRM and PEMS for the (a) NO<sub>x</sub> (b)CO<sub>2</sub> (c)THC and (d) CO

## **4.5 Data Analysis Protocol within the NTE Zone**

The characterization of emission measurements during NTE events was a critical component of the chassis dynamometer testing and analysis. For this task, two cycles were specifically designed to simulate operation in the NTE zone and transitions into and out of the NTE zone. These two were the NTE steady state cycles and the NTE stepped cycle. NTE events were also identified in the inventory cycles, namely the UDDS and the 50 mph cruise.

This section simply reports on comparisons between the PEMS2 and PEMS4 and the FRM for events in the NTE zone. The NTE analysis primarily focused on NTE events that were determined based on information obtained from the engine control module data and the J1939 signal. For PEMS4, the engine data connection did not work, so the NTE events for PEMS4 were determined based on the data collected by UCR. The FRM sampled the J1939 signal at 10 Hz and recorded every  $10^{th}$  point for a discrete 1 Hz signal (with no averaging) for NTE determination. PEMS2 sampled at the maximum J1939 broadcast data rate (up to 10 Hz depending on the parameter), averaged the data over 1/4 second intervals, and recorded data at that rate. The data were then interpolated between the 4 Hz averages to the precise 1 second reporting intervals. Since PEMS4 used the FRM data, subsequent comparisons of identified NTE events (e.g., Table 4-2 below) and bhp-hr (e.g., Figure 4-8) only include the FRM and PEMS2.

#### **4.6 Measurements of Brake-Horsepower Hour**

Emission certification standards are reported in emissions per unit or work, thus measurement of work is an important issue for compliance testing. The work recorded within the NTE zone was based on measurements obtained from the J1939 signal and integrated over the individual NTE events to determine the brake horsepower hour for each NTE event. Figure 4-8 shows the correlation between the bhp calculated by the FRM and PEMS2. The correlation plot represents a filtered set of data in which only NTE events that "matched" between the FRM and PEMS2 are included and for the entire set of NTE data. In this case, matching NTE events are ones where the identified NTE events has a similar start time between the FRM and PEMS2 and a duration that is the same within 4 seconds. Overall, bhp showed good correlation on an NTE event basis, with the outlier data primarily due to different identification of a specific NTE event. Differences in the identification of specific NTE events are discussed further below.



**Figure 4-8 Correlation Plot for bhp for Individual NTE Events** 

### **4.7 Identifying the NTE Events in the Steady State Cycle**

The NTE steady state cycles were designed to have steady state operation with transitions into and out of the NTE zone by changing load. Each test segment included seven replicates with identifiable NTE events. To help in the identification of the NTE event we created an ID number that included the engine speed, followed by the iteration, followed by the number of the NTE event as it was identified within the cycle. For example, the first event was identified as NTE\_1290\_1\_1. The load is also presented in a separate column.

value for the NTE onset after subtracting the frictional torque. In general, the number and duration of NTE events was relatively consistent for the FRM and PEMS. There were typically 21 NTE events within a cycle at a particular engine speed, corresponding to 7 iterations at 3 engine load points. The 1290 engine speed/40% load was generally outside the NTE zone because the torque was below the threshold

A breakdown of the NTE events is shown in Figure 4-9 and a comparison of rpm and torque signals for the FRM and PEMS2 is provided in Figures 4-10. For PEMS4, emissions data were collected for only one of the two cycles run at the different speeds and the final NTE event was not identified in the PEMS4 analysis. The FRM also only collected two data points for  $NO<sub>x</sub>$  for the first run at 1550 at the highest load point due to an electronic problem with the trigger to the PEMS instruments.



**NTE Steady State Cycle - 1770 RPM Run #1** 

**Figure 4-9. Summary of NTE Events identified during the NTE Steady State Cycle.** 



**Figure 4-10. Comparison of Torque and RPM Signals for Typical NTE Event** 

A comparison of each identified NTE event is provided in Table 4-2, including the start time and the duration determined by the FRM and the PEMS.

Unique ID for NTE	FRM	PEMS2	PEMS <sub>2</sub>	<b>FRM</b>	PEMS2	PEMS <sub>2</sub>
event			NTEstart NTEstart difference	<b>NTEdur</b>		NTEdur difference
NTE_1290_1_1	389	389	0	40	40	0
NTE 1290 1 2	439	439	0	40	40	0
NTE 1290 1 3	489	489	0	40	40	0
NTE 1290 1 4	539	539	0	40	40	0
NTE 1290 1 5	589	589	0	40	40	0
NTE 1290 1 6	639	639	0	41	41	0
NTE 1290 1 7	689	689	0	40	40	0
NTE 1290 1 8	759	759	0	40	40	0
NTE 1290 1 9	809	809	0	40	40	0
NTE_1290_1_10	859	859	0	40	40	0
NTE_1290_1_11	909	909	0	40	40	0
NTE_1290_1_12	959	959	0	40	40	0
NTE 1290 1 13	1009	1009	0	40	40	0
NTE 1290 1 14	1059	1059	0	40	40	0
NTE 1290 2 1	391	391	0	40	40	0
NTE 1290 2 2	441	441	0	40	40	0
NTE 1290 2 3	491	491	0	40	40	0
NTE 1290 2 4	541	541	0	40	40	0
NTE 1290 2 5	591	591	0	40	40	$\Omega$

**Table 4-2. Identified NTE Events for the FRM and PEMS over the NTE Steady State Cycle** 





### **4.8 Analysis of the Data for the Steady-state NTE Cycle**

Before discussing into the full analysis, the approach used needs to be explained.

- 1. Results are expressed as g/bhp-hr like certification standards and on a fuel specific (fs) basis. The fs calculations were based on fuel use determined from  $CO<sub>2</sub>$  emissions measurements so in part will normalize flow differences between the different instruments.
- 2. Results are expressed relative to the approximate NTE thresholds for 2007 model year engines since a determination of the error margin for PEMS units relative to the NTE standard is an important consideration.

The full results for the individual tests are provided in Appendix H on a g/bhp-hr and fuel specific (fs) basis. Statistical comparisons of the NTE steady state results are provided in Appendix I. The statistical analyses included a paired and unpaired t-test and an F-test for comparison of experimental variance. The statistical comparisons for the steady state cycle showed that in most cases the differences between the mass emission rates for the different instruments were statistically significant at greater than a 95% confidence level. The statistical significance of the F-test results depended on the specific test iteration and the specific emission component.

The average NTE results in  $g/b$ hp-hr and in fuel specific units for NO<sub>x</sub> over the steady state cycles are provided in Figure 4-11a and 4-11b.



**Figure 4-11 Comparison of NOx Emission Rates and 95% Confidence Limits. (a) g/bhp-hr, (b) fsNOx** 





Figure 4-12. Correlation Plot for g/bhp-hr NO<sub>x</sub> emissions over all NTE Steady State Cycle **Events** 

As evident in Figure 4-12, the correlation plots show reasonable correlation between the FRM and PEMS  $(R^2 = 0.91$  and 0.92), with each PEMS showing a bias from the FRM. Clearly, PEMS 2 is much closer to the values measured by the FRMs.

Separate plots of the percentage difference for the PEMS in comparison with the FRM are provided in Figure 4-13a and 4-13b, respectively, for g/bhp-hr relative to the NTE thresholds and fs units. The absolute values of the differences are provided in Table 4-3.



Figure 4-13. Comparison of Percentage Differences in NO<sub>x</sub> Emission Rates for Different **PEMS. a) g/bhp-hr, b) fsNOx** 

The percentage differences for the NTE events were calculated relative to the NTE thresholds, since these are the differences most relevant for regulatory implementation. For  $NO_x$ , this was done by taking the absolute difference in the mass emission rates for the FRM and a particular PEMS, dividing by the NTE threshold  $(2.0 \text{ g/bhp-hr}$  for NO<sub>x</sub>), and presenting the value as a percent. The  $NO<sub>x</sub>$  emission rates for PEMS2 was approximately 3 to 17% higher than those for the FRM in g/bhp-hr units relative to the NOx NTE threshold. PEMS 4 showed larger differences compared to the FRM, with percent differences ranging from 26 to 47%. These larger differences could be due in part to the calibration offset and associated software issues for PEMS4, as noted above. On a fs basis, PEMS2 showed slightly lower differences with the FRM ranging from 2-11%. The comparison of fs  $NO<sub>x</sub>$  emissions improved for PEMS4, with differences ranging from 9-18%. This is probably due to the fact that both the  $NO<sub>x</sub>$  and  $CO<sub>2</sub>$  calibrations for PEMS4 were relatively high, which would be offset when the emissions are ratioed.

The average THC NTE results in g/bhp-hr and the percentage difference relative to the NTE threshold for the PEMS in comparison with the FRM are provided in Figure 4-14a and 4-14b, respectively. THC was generally lower for both PEMS2 and PEMS4 compared with the FRM. THC emissions were below the NTE threshold standards for these measurements. Relative to the approximate NTE threshold for THC of 0.21 g/bhphr, the percentage differences were all within 10%, with most being within 5%.



**Figure 4-14. (a) Comparison of Mass Emission Rates for THC in g/bhp-hr (b) Percentage Difference of NTE Standard between PEMS and FRM for THC in g/bhp-hr** 

The correlation plot for THC g/bhp-hr emissions is provided in Figure 4-15. The correlation coefficient for THC emissions was not as good as for  $NO_x$  emissions as  $R^2 =$ 0.78 and 0.74, respectively, for PEMS2 and PEMS4.



**Figure 4-15. Correlation Plot for g/bhp-hr THC Emissions over all NTE Steady State Cycle Events** 

The average CO NTE results in g/bhp-hr and the percentage difference for the PEMS in comparison with the FRM are provided in Figure 4-16a and 4-16b, respectively. CO emissions for PEMS2 were consistently higher than those for the FRM. However, the CO emissions were well below the NTE threshold of 19.4 g/bhp-hr, and the percentage differences between the FRM and PEMS2 were all within 2% relative to the NTE CO threshold.



**Figure 4-16. (a) Comparison of Mass Emission Rates for CO in g/bhp-hr (b) Percentage Difference of NTE Standard between PEMS and FRM for CO in g/bhp-hr** 

The correlation plot over the whole range of measured values is depicted in Figure 4-17 and shows an  $\mathbb{R}^2$ =0.99, with a definitive bias; PEMS 2 is always higher than the FRM values.



**Figure 4-17 Correlation Plot for CO g/bhp-hr Emissions over all NTE Steady State Cycle Events** 

The average  $CO<sub>2</sub> NTE$  results in g/bhp-hr and the percentage difference for the PEMS in comparison with the FRM are provided in Figure 4-18a and 4-18b, respectively. Results depended on the PEMS unit. Results from PEMS2 were generally 5% higher than the FRM while results from PEMS4 varied from 10 to 30% higher.



Figure 4-18. (a) Comparison of Mass Emission Rates for CO<sub>2</sub> in g/bhp-hr (b) Percentage Difference between PEMS and FRM for CO<sub>2</sub> in g/bhp-hr

The correlation plot for  $CO<sub>2</sub>$  is shown in Figure 4-19. PEMS2 showed good correlation with the FRM ( $\mathbb{R}^2 = 0.99$ ), but a slight bias. PEMS4 CO<sub>2</sub> emissions were 9 to 30% higher than those the FRM, with a relatively poor correlation coefficient ( $R^2 = 0.57$ ). The

magnitude of the bias for PEMS4 is in the range of the calibration offset for PEMS4, although the poor correlation indicates that other factors also contribute to this difference.



Figure 4-19. Correlation Plot for CO<sub>2</sub> g/bhp-hr Emissions over all NTE Steady State Cycle **Events** 

Table 4-3 below provides the total stream used in the charts and analysis that were just reviewed and included for completeness.



**Table 4-3. Absolute Differences in g/bhp-hr Emissions for the Steady State Cycle NTE Events** 

#### **4.9 Analysis of the Data for the NTE Stepped Cycle**

This cycle consisted of a series of speeds/loads that were stepwise increased and then decreased. The number of NTE events identified for this varied from 12 to 15, depending on the test iteration and the measurement device. Figure 4-20 shows graphically when the NTE events occurred within the cycle. Twelve primary events were built into the experimental design and are the focus of the subsequent analysis. These are labeled 1-12 in Figure 4-20. Some "miscellaneous" NTE events were also identified intermittently for a subset of tests by one or more of the instruments. These additional miscellaneous NTE events are labeled "m1-m3" in the Figure. A summary of the NTE events for all tests is provided in Table 4-4. It must be noted that the data for PEMS4 was only available for the final three stepped cycles. PEMS4 also did not identify any NTE events that occurred after 1,000 seconds into the cycle.



**NTE Stepped Cycle Run #1** 

**Figure 4-20. Summary of NTE Events for the NTE Stepped Cycle** 

	<b>FRM</b>	PEMS <sub>2</sub>	PEMS <sub>2</sub>	<b>FRM</b>	PEMS <sub>2</sub>	PEMS <sub>2</sub>
NTE event			NTEstart NTEstart difference			NTEdur NTEdur difference
#1						
NTE_Stepped_1_1	112	113	$\mathbf{1}$	95	95	0
NTE_Stepped_2_1	102	#N/A	#N/A	100	#N/A	#N/A
NTE_Stepped_3_1	106	106	0	96	97	1
NTE_Stepped_4_1	124	124	0	97	97	0
NTE_Stepped_5_2	105	105	0	101	102	1
NTE_Stepped_6_1 #2	106	106	0	99	99	0
NTE_Stepped_1_2	209	209	0	100	101	$\mathbf{1}$
NTE_Stepped_2_2	204	204	0	114	114	0
NTE_Stepped_3_2	204	205	1	110	110	0
NTE_Stepped_4_2	224	224	0	112	112	0
NTE_Stepped_5_3	209	209	0	105	105	0
NTE_Stepped_6_2	207	207	0	117	117	0
#3						
NTE_Stepped_1_3	311	311	0	59	60	1
	319	319	0	44	44	0
NTE_Stepped_2_3					44	
NTE_Stepped_3_3	318	316	$-2$	42		$\overline{c}$
NTE_Stepped_4_3	338	338	0	47	47	0
NTE_Stepped_5_4	316	317	1	43	43	0
NTE_Stepped_6_3	#N/A	326	#N/A	#N/A	40	#N/A
#4						
NTE_Stepped_1_4	384	385	1	74	74	0
NTE_Stepped_2_4	377	377	0	75	74	$-1$
NTE_Stepped_3_4	372	373	1	80	79	$-1$
NTE_Stepped_4_4	399	399	0	72	72	0
NTE_Stepped_5_5	373	373	0	76	76	0
NTE_Stepped_6_4	374	374	0	78	78	0
#5						
NTE_Stepped_1_5	470	471	1	79	78	-1
NTE_Stepped_2_5	461	461	0	80	80	0
NTE_Stepped_3_5	460	461	1	81	81	0
NTE_Stepped_4_5	481	482	1	81	80	-1
NTE_Stepped_5_6	458	458	0	82	82	0
NTE_Stepped_6_5	462	462	0	80	80	0
#6						
NTE_Stepped_1_6	567	568	1	50	51	1
NTE_Stepped_2_6	557	557	0	53	53	0
NTE_Stepped_3_6	556	557	1	54	54	0
NTE_Stepped_4_6	580	580	0	51	51	0
NTE_Stepped_5_7	557	558	1	53	36	$-17$
NTE_Stepped_6_6	560	560	0	66	66	0
#7						
NTE_Stepped_1_7	829	829	0	99	99	0

**Table 4-4. NTE Event Summary for NTE Stepped Cycle** 



**Note**: 114 seconds added to NTE Start for PEMS2 on test iteration #2 because the PEMS2 sampler was started late for this cycle, due to the late start NTE event #1 is also missing for PEMS2 for this test iteration.

The average results for the NTE stepped cycle are presented in Table 4-5 on a g/bhp-hr and fuel specific basis. The results are presented in chronological order for the 12 major NTE events within the cycle. The average values are derived based on the results of the 6 test iterations for the specific NTE event. The statistical comparisons of the differences in emissions between the FRM and PEMS are provided in Appendix J. Generally, most differences observed between the FRM and PEMS were statistically significant at greater than the 95% confidence limit.

The general trends for emissions for the stepped cycle are similar to those found for the NTE steady state cycle and the integrated results. On a g/bhp-hr basis,  $NO<sub>x</sub>$  emissions for PEMS2 ranged from 6 to 11% higher than the FRM relative to the NTE threshold, while  $NO<sub>x</sub>$  emissions for PEMS4 were from 17 to 48% higher than the FRM relative to the NTE threshold.

For  $CO<sub>2</sub>$ , PEMS2 ranged from 3.2% to 6.5% higher than the FRM, while PEMS4 was 19 to 25% higher than the FRM. For PEMS4, some of these differences can be attributed to the offset for the calibration. CO also showed a consistent positive bias for PEMS2 relative to the FRM for CO, but this represented less than a 1% difference relative to the NTE threshold. THC emissions as a whole were relatively low. THC emissions for PEMS2 tended to the lower than those for the FRM, while THC emissions for PEMS4 were mixed relative to the FRM. For both PEMS, the percentage differences represented less than 5% of the NTE standard for all NTE events.





**Table 4-5 Comparison of Average g/bhp-hr and fs Emissions for FRM and PEMS for the NTE Stepped Cycle** 

#### **4.10 Analysis of the Data for the 50 mph cruise**

The 50 mph cruise had between 3 to 6 NTE events over the 7 replicate runs. Four primary NTE events were consistently identified over the replicate test runs. The NTE events are labeled 1-4 in Figure 4-21. Other NTE events varied throughout the course of the cruise portion of the cycle. These additional NTE events that were identified for at least one of the test iterations are labeled "m" in Figure 4-21. As an example, Figure 4-22 shows the torque, RPM and engine load for one test iteration where differences in identified NTE events were observed between the FRM and PEMS2.



**Figure 4-21. Summary of the NTE Events Identified During the 50 mph Cruise Cycle** 



**Figure 4-22. Comparison of Torque and RPM Signals for Unmatched NTE Events** 

Table 4-6 shows the number of NTE events identified by the FRM and PEMS2 for each test run.

	<b>FRM</b>	PEMS2	PEMS <sub>2</sub>	<b>FRM</b>	PEMS2	PEMS <sub>2</sub>
NTE event			NTEstart NTEstart difference NTEdur NTEdur difference			
#1						
50CRUISE 1 1	186	186	$\overline{0}$	91	309	218
50CRUISE 2 1	223	192	$-31$	34	65	31
50CRUISE 3 1	187	188	$\mathbf{1}$	38	34	$-4$
50CRUISE 4 1	192	126	$-66$	66	66	$\boldsymbol{0}$
50CRUISE 5 1	193	194	$\mathbf{1}$	68	67	$-1$
50CRUISE 6 1	194	195	$\mathbf{1}$	65	64	$-1$
50CRUISE 7 1	191	192	$\mathbf{1}$	66	65	$-1$
#2						
50CRUISE 1 2	278	$\#N/A$	$\sharp N/A$	217	$\sharp N/A$	$\sharp N/A$
50CRUISE 2 2	263	263	$\boldsymbol{0}$	225	282	57
50CRUISE 3 3	265	265	$\boldsymbol{0}$	219	154	$-65$
50CRUISE 4 2	262	197	$-65$	283	282	$-1$
50CRUISE 5 2	266	267	$\mathbf{1}$	54	54	$\overline{0}$
50CRUISE 6 2	265	265	$\boldsymbol{0}$	283	284	1
50CRUISE 7 2	260	262	$\overline{2}$	283	211	$-72$
#3						
50CRUISE 1 3	496	496	$\mathbf{0}$	47	48	$\mathbf{1}$
50CRUISE 2 3	489	$\sharp N/A$	$\sharp N/A$	56	$\sharp N/A$	$\sharp N/A$
50CRUISE 3 5	$\sharp N/A$	496	$\sharp N/A$	$\sharp N/A$	47	$\sharp N/A$
50CRUISE 5 5	496	497	$\mathbf{1}$	78	78	$\overline{0}$
50CRUISE 7 3	$\sharp N/A$	474	$\sharp N/A$	$\sharp N/A$	70	$\sharp N/A$
#4						
50CRUISE 1 4	547	548	$\mathbf{1}$	86	119	33
50CRUISE 2 4	550	550	$\boldsymbol{0}$	134	134	$\overline{0}$
50CRUISE 3 6	549	549	$\mathbf{0}$	120	121	$\mathbf{1}$
50CRUISE 4 3	547	482	$-65$	123	123	$\boldsymbol{0}$
50CRUISE 6 3	550	550	$\boldsymbol{0}$	134	136	$\overline{c}$
50CRUISE 7 4	548	549	$\mathbf{1}$	120	120	$\overline{0}$
misc.						
50CRUISE 3 2	226	223	$-3$	34	38	$\overline{4}$
50CRUISE 5 3	322	322	$\overline{0}$	60	61	$\mathbf{1}$
50CRUISE 5 4	383	384	1	106	106	$\boldsymbol{0}$
50CRUISE 3 4	$\sharp N/A$	420	$\sharp N/A$	$\sharp N/A$	67	$\sharp N/A$
50CRUISE 5 6	575	592	17	94	78	$-16$

**Table 4-6. NTE Event Summary for the 50 mph Cruise** 

The results for the g/bhp-hr and fuel-specific emissions for the 50 mph Cruise cycle are presented in Table 4-7 and 4-8. The percentage differences for the brake specific  $NO<sub>x</sub>$ , CO, and THC are given relative to the applicable NTE threshold. The Tables include only PEMS2, since data for PEMS4 are available for only a single test cycle iteration. The results are generally consistent with the other cycles. The percentage difference from the FRM relative to the NTE standard for  $NO_x$  was between –1 to 12% for PEMS2 and 3672% for PEMS4. For THC, the percentage differences represented 5% or less of the NTE threshold, while CO measurement differed by  $1\%$  or less of the NTE threshold.  $CO<sub>2</sub>$ measurements for PEMS2 were within 7% of the FRM for all NTE events, while PEMS4 showed larger differences of 22-39%. The differences between the PEMS and the FRM were generally statistically significant over the data set.

NTE event	<b>FRM</b> bsNOx		PEMS2 PEMS2 vs. FRM bsNOx absol. Diff vs. NTE	PEMS2 % diff	<b>FRM</b> bsTHC	PEMS2 bsTHC	PEMS2 Diff	vs. FRM PEMS2 absol. % diff vs. NTE	<b>FRM</b> bsCO	PEMS <sub>2</sub> bsCO	PEMS2 absol. Diff	vs. FRM PEMS2 % diff vs. NTE	<b>FRM</b> bsCO <sub>2</sub>	PEMS <sub>2</sub> bsCO <sub>2</sub>	PEMS2 absol. Diff	vs. FRM PEMS2 % diff vs. FRM
50CRUISE_1_1	2.27	2.25	$-0.02$	$-1%$	0.020	0.023	0.00	2%	0.46	0.61	0.15	0.8%	538	573	34.83	6%
50CRUISE_1_2	2.08	#N/A	#N/A	#N/A	0.024	#N/A	#N/A	#N/A	0.56	#N/A	#N/A	#N/A	570	#N/A	#N/A	#N/A
50CRUISE_1_3	2.08	2.19	0.11	5%	0.024	0.022	0.00	$-1%$	0.53	0.59	0.06	0.3%	561	582	20.41	4%
50CRUISE_1_4	2.07	2.19	0.12	6%	0.026	0.024	0.00	$-1%$	0.54	0.61	0.07	0.4%	563	589	25.98	5%
50CRUISE_2_1	2.15	2.43	0.28	14%	0.016	0.014	0.00	$-1%$	0.49	0.53	0.04	0.2%	571	529	$-41.88$	$-7%$
50CRUISE_2_2	2.11	2.25	0.14	7%	0.016	0.018	0.00	1%	0.51	0.64	0.13	0.7%	572	583	11.07	2%
50CRUISE_2_3	2.10	#N/A	#N/A	#N/A	0.015	#N/A	#N/A	#N/A	0.49	#N/A	#N/A	#N/A	567	#N/A	#N/A	#N/A
50CRUISE_2_4	2.13	2.28	0.15	8%	0.016	0.018	0.00	1%	0.48	0.65	0.17	0.9%	570	586	15.55	3%
50CRUISE_3_1	2.33	2.44	0.11	5%	0.017	0.015	0.00	$-1%$	0.38	0.45	0.07	0.4%	528	518	$-10.45$	$-2%$
50CRUISE 3 2	2.22	2.37	0.15	7%	0.026	0.030	0.00	2%	0.50	0.68	0.19	1.0%	582	585	3.37	1%
50CRUISE_3_3	2.15	2.27	0.12	6%	0.025	0.019	$-0.01$	$-3%$	0.49	0.66	0.18	0.9%	571	576	4.89	1%
50CRUISE_3_4	#N/A	2.27	#N/A	#N/A	#N/A	0.017	#N/A	#N/A	#N/A	0.63	#N/A	#N/A	#N/A	594	#N/A	#N/A
50CRUISE_3_5	#N/A	2.25	#N/A	#N/A	#N/A	0.014	#N/A	#N/A	#N/A	0.61	#N/A	#N/A	#N/A	572	#N/A	#N/A
50CRUISE_3_6	2.20	2.35	0.15	8%	0.026	0.016	$-0.01$	$-5%$	0.45	0.61	0.16	0.8%	574	587	13.56	2%
50CRUISE_4_1	2.31	2.51	0.20	10%	0.018	0.012	$-0.01$	$-3%$	0.44	0.56	0.12	0.6%	535	534	$-0.60$	0%
50CRUISE_4_2	2.17	2.35	0.18	9%	0.022	0.016	$-0.01$	$-3%$	0.45	0.62	0.17	0.9%	572	586	14.19	2%
50CRUISE_4_3	2.19	2.34	0.15	7%	0.022	0.016	$-0.01$	$-3%$	0.44	0.60	0.16	0.8%	572	588	15.73	3%
50CRUISE_5_1	2.31	2.40	0.09	4%	0.017	0.022	0.01	3%	0.43	0.46	0.03	0.2%	544	554	9.51	2%
50CRUISE_5_2	2.19	2.29	0.10	5%	0.018	0.027	0.01	4%	0.46	0.53	0.07	0.4%	543	576	32.59	6%
50CRUISE_5_3	2.19	2.30	0.11	6%	0.019	0.028	0.01	4%	0.47	0.53	0.07	0.3%	554	583	28.92	5%
50CRUISE 5 4	2.18	2.27	0.09	4%	0.019	0.029	0.01	5%	0.43	0.51	0.08	0.4%	567	595	28.17	5%
50CRUISE_5_5	2.18	2.26	0.08	4%	0.019	0.029	0.01	5%	0.44	0.52	0.08	0.4%	565	592	27.16	5%
50CRUISE_5_6	2.21	2.29	0.08	4%	0.020	0.028	0.01	4%	0.42	0.48	0.07	0.3%	564	591	27.12	5%
50CRUISE 6 1	2.34	2.47	0.13	7%	0.017	0.020	0.00	2%	0.43	0.52	0.09	0.5%	532	537	5.45	1%
50CRUISE_6_2	2.21	2.29	0.08	4%	0.020	0.028	0.01	4%	0.43	0.54	0.11	0.6%	564	590	25.96	5%
50CRUISE_6_3	2.22	2.30	0.08	4%	0.020	0.028	0.01	4%	0.42	0.55	0.12	0.6%	560	590	29.70	5%
50CRUISE_7_1	2.24	2.48	0.24	12%	0.023	0.021	0.00	$-1%$	0.53	0.64	0.10	0.5%	514	543	28.70	6%
50CRUISE_7_2	1.95	2.16	0.21	11%	0.028	0.028	0.00	0%	0.65	0.80	0.15	0.8%	560	589	28.60	5%
50CRUISE_7_3	#N/A	2.15	#N/A	#N/A	#N/A	0.027	#N/A	#N/A	#N/A	0.78	#N/A	#N/A	#N/A	597	#N/A	#N/A
50CRUISE_7_4	1.94	2.16	0.22	11%	0.028	0.027	0.00	0%	0.64	0.81	0.17	0.9%	557	591	33.96	6%

**Table 4-7. g/bhp-hr NTE Results for the 50 mph Cruise Cycle**
ID Unique for		PEMS <sub>2</sub>	PEMS <sub>2</sub>		PEMS <sub>2</sub>	PEMS2		PEMS <sub>2</sub>	PEMS <sub>2</sub>
NTE event	<b>FRM fsNOx</b>	fsNOx	% diff	<b>FRM fsTHC</b>	fsTHC	% diff	<b>FRM fsCO</b>	fsCO	% diff
50CRUISE_1_1	0.0133	0.0124	$-7%$	0.000117	0.000133	14%	0.0027	0.0035	28%
50CRUISE_1_2	0.0115	#N/A	#N/A	0.000135	#N/A	#N/A	0.0031	#N/A	#N/A
50CRUISE 1 3	0.0117	0.0120	2%	0.000132	0.000122	$-8%$	0.0030	0.0033	11%
50CRUISE_1_4	0.0116	0.0119	3%	0.000143	0.000131	$-9%$	0.0030	0.0033	10%
50CRUISE 2_1	0.0119	0.0144	21%	0.000088	0.000086	$-2%$	0.0027	0.0032	20%
50CRUISE_2_2	0.0117	0.0124	6%	0.000088	0.000099	13%	0.0028	0.0035	26%
50CRUISE 2 3	0.0117	#N/A	#N/A	0.000084	#N/A	#N/A	0.0027	#N/A	#N/A
50CRUISE_2_4	0.0118	0.0125	6%	0.000086	0.000097	12%	0.0027	0.0036	34%
50CRUISE 3 1	0.0140	0.0150	8%	0.000103	0.000096	$-6%$	0.0023	0.0028	23%
50CRUISE_3_2	0.0121	0.0130	8%	0.000141	0.000164	16%	0.0027	0.0038	40%
50CRUISE 3 3	0.0119	0.0126	6%	0.000139	0.000107	$-23%$	0.0027	0.0037	38%
50CRUISE_3_4		0.0122	#N/A	#N/A	0.000093	#N/A	#N/A	0.0034	#N/A
50CRUISE 3 5		0.0126	#N/A	#N/A	0.000082	#N/A	#N/A	0.0035	#N/A
50CRUISE_3_6 <sup>#N/A</sup> 0.0121		0.0128	6%	0.000142	0.000086	$-39%$	0.0025	0.0033	35%
50CRUISE $4.1^{#N/A}$ 0.0136		0.0148	9%	0.000104	0.000074	$-29%$	0.0026	0.0033	30%
50CRUISE_4_2	0.0120	0.0128	7%	0.000120	0.000087	$-28%$	0.0025	0.0034	36%
50CRUISE 4 3	0.0121	0.0128	6%	0.000121	0.000086	$-29%$	0.0024	0.0033	36%
50CRUISE_5_1	0.0134	0.0137	2%	0.000096	0.000135	41%	0.0025	0.0027	9%
50CRUISE 5 2	0.0127	0.0127	0%	0.000107	0.000149	39%	0.0026	0.0030	11%
50CRUISE_5_3	0.0125	0.0127	2%	0.000109	0.000156	43%	0.0027	0.0030	12%
50CRUISE 5 4	0.0122	0.0122	0%	0.000108	0.000156	45%	0.0024	0.0028	14%
50CRUISE_5_5	0.0122	0.0123	1%	0.000109	0.000158	45%	0.0025	0.0029	15%
50CRUISE_5_6	0.0124	0.0124	0%	0.000111	0.000154	39%	0.0023	0.0026	12%
50CRUISE 6_1	0.0139	0.0145	4%	0.000098	0.000123	25%	0.0026	0.0031	21%
50CRUISE 6 2	0.0124	0.0125	1%	0.000112	0.000150	34%	0.0024	0.0029	23%
50CRUISE_6_3	0.0125	0.0125	0%	0.000113	0.000152	35%	0.0024	0.0030	25%
50CRUISE_7_1	0.0137	0.0142	3%	0.000139	0.000131	$-6%$	0.0033	0.0038	17%
50CRUISE_7_2	0.0110	0.0117	7%	0.000159	0.000155	$-3%$	0.0037	0.0044	20%
50CRUISE 7 3		0.0115	#N/A	#N/A	0.000148	#N/A	#N/A	0.0042	#N/A
50CRUISE 7 4	0.0110	0.0117	6%	0.000156	0.000144	$-8%$	0.0036	0.0044	21%
	#N/A								

 **Table 4-8. Fuel-Specific NTE Results for the 50 mph Cruise Cycle** 

#### **4.11 Analysis of the Data for the Urban Dynamometer Driving Schedule (UDDS)**

Between one and three NTE events were identified for the UDDS over the 7 test iterations. The primary NTE event for this cycle occurred approximately 632 seconds into the cycle as the vehicle is approaching its highest speed point, as shown in Figure 4- 23. This NTE event typically lasted between 30-35 seconds. One or two other NTE events were identified for most of the UDDS cycles, but these were not uniformly identified by all PEMS/FRM. A summary of the identified NTE events identified by the FRM and both PEMS for each test run is provided in Table 4-9. For all identified NTE events, the duration of the event was essentially the same for all PEMS/FRM.



**Figure 4-23. Summary of the NTE Events Identified during the UDDS Cycle** 

	<b>FRM</b>	PEMS <sub>2</sub>	PEMS <sub>2</sub> NTEstart NTEstart % difference NTEdur	<b>FRM</b>	PEMS <sub>2</sub> <b>NTEdur</b>	PEMS <sub>2</sub> %differencel
UDDS MTA 1 1	632	632	0	34	34	0
UDDS MTA 2 1	632	632	0	33	33	0
UDDS MTA 2 2	667	668	-1	32	32	0
UDDS MTA 2 3	702	703	-1	38	38	0
UDDS MTA 3 1	632	632	0	34	33	
UDDS MTA 3 2	705	707	$-2$	37	35	2
UDDS MTA 4 1	NA.	594	NА	ΝA	34	NА
UDDS MTA 4 2	634	633	1	35	30	5
UDDS MTA 4 3	NA.	NA.	NА	ΝA	NA.	NА
UDDS MTA 5 1	633	633	0	34	34	0
UDDS MTA 5 2	NA.	669	NА	ΝA	32	NA.
UDDS MTA 6 1	635	634	1	33	34	-1
UDDS MTA 7 1	635	636	-1	34	33	1
UDDS MTA 7 2	NА	672	ΝA	ΝA	31	ΝA

**Table 4-9. NTE Event Comparison for UDDS Cycle** 

Tables 4-10 and 4-11 show emissions for each individual NTE event in g/bhp-hr and g/grams of fuel, respectively, along with the associated percentage differences with the FRM. The comparisons between the FRM and PEMS for the different NTE events are similar to those found for the other cycles and were statistically significant in most cases.



# **Table 4-10. Comparison of Average g/bhp-hr Emissions for FRM and PEMS for the UDDS Cycle**

								THC					CO		
		P <sub>2</sub>	P <sub>4</sub>	P <sub>2</sub>	P <sub>4</sub>	<b>FRM</b>	P <sub>2</sub>	P4		P4	<b>FRM</b>	P <sub>2</sub>	P <sub>4</sub>	P <sub>2</sub>	P4
					% diff					% diff					% diff
					vs. FRM		Fs	P <sub>2</sub>		vs. FRM		fs			vs. FRM
UDDS_MTA_1_1	0.0141	0.0147	0.0143	4%	1%			0.00208	4%	1249%	0.0035	0.0048	NA	36%	<b>NA</b>
UDDS_MFRAM2_1	.0142		0.0144 0.0146	1%	3%				68%	57%		0.0034 0.0041	ΝA	21%	<b>NA</b>
UDDS_MTA_2_2	.0151		0.0158 0.0165	5%	9%				70%	69%		0.0035 0.0045	NA	26%	<b>NA</b>
UDDS MTA 2 3	.0148		0.0150 0.0154	1%	4%				71%	71%	0.0037	0.0042	NA	15%	<b>NA</b>
UDDS_MTA_3_	.0139		0.0148 0.0154	6%	11%			0.00013	$-19%$	-6%	0.0031	0.0038	ΝA	24%	<b>NA</b>
UDDS_MMAx <sup>3_</sup> $\overline{2}$			0.0153 0.0163	4%	10%		0.00012 0.00010	0.00012	$-20%$	-2%	0.0033	0.0041	NA	25%	<b>NA</b>
UDDS_MJA_4_1	0.0148 NA	0.0147	NA.	<b>NA</b>	ΝA	ΝA	0.00019	<b>NA</b>	<b>NA</b>	<b>NA</b>	NA	0.0040	NA	NA	<b>NA</b>
UDDS_MTA_4_2			0.0162 0.0150	15%	7%		0.00007 0.00019	0.00016	175%	124%	0.0031	0.0041	NA	33%	<b>NA</b>
UDDS MTA 4 3	.0141 NA	NA.	0.0166	<b>NA</b>	<b>NA</b>	ΝA	NA	0.00016	<b>NA</b>	<b>NA</b>	NA	NA	<b>NA</b>	NA	<b>NA</b>
UDDS_MTA_5_1			0.0152 0.0154	5%	7%			0.00013	57%	51%	0.0031	0.0042	NA	36%	<b>NA</b>
UDDS MTA 5 2	0144. NA	0.0163	NА	<b>NA</b>	<b>NA</b>	NA	0.00014	ΝA	<b>NA</b>	<b>NA</b>	NA	0.0044	NA	NA	<b>NA</b>
UDDS_MTA_6_1			0.0146 0.0146	7%	7%		0.00014 0.00013	0.00016	$-10%$	8%	0.0037	0.0047	NA	27%	<b>NA</b>
UDDS_MTA_7_	.0136		0.0149 0.0146	10%	8%		0.00012 0.00017	0.00015	45%	31%	0.0039	0.0050	ΝA	28%	<b>NA</b>
UDDS_MTA_7_2	.0135 NA	0.0150	NA	ΝA	ΝA	ΝA	0.00017	ΝA	NA	ΝA	ΝA	0.0052	NA	<b>NA</b>	<b>NA</b>

**Table 4-11 Comparison of Average fs Emissions for FRM and PEMS for the UDDS Cycle** 

# **5.0 Exploratory Survey of PM-Capable PEMS for Mobile Sources**

## **5.1 Experimental Procedures**

The earlier section on measuring emissions from stationary sources presented some comparative data for PEMS and the FRM. By the time of the mobile source testing many more PM-capable PEMS were available so the thrust of this section is to describe the comparison between the FRM and four different PM-capable PEMS during the chassis dynamometer testing portion of this research. One PEMS used conventional filter methods and three used real-time second-by-second mass concentration methods. Table 5-1 and 5-2 lists the principle of operation and specifications for each of the PM measurement techniques. Additionally, PEMS6 for measuring PM was included in the study for the chassis dynamometer testing. After the testing, it was determined that the instrument had not been operating within manufacturers' specifications, so these data were not reported. Note that the PM-capable PEMS are an evolving product, like the gaseous PEMS, so there is little standardization. Diversity in manufacturer approach is expected.

UCR's MEL took PM mass samples with conventional Gelman Teflo® 47 mm filters and the elemental and organic carbon (EC/OC) samples on specially prepared quartz media. Following the CFR protocol, all filter media were sampled using secondary dilution, a 47°C filter temperature and an inlet classifier as per CFR for 2007. The secondary dilution tunnel was operated at a dilution ratio of 2.5/1, and a filter sample flow rate of 19 standard liters per minute (slpm), giving filter face velocities of around 37 cm/s for both the quartz and Teflon filters. The NIOSH method was used to measure the EC and OC and to estimate the mass. PEMS 5, 7, and 8 all sampled from the FRM primary dilution tunnel. PEMS 3 sampled from its own dilution tunnel connected to the raw exhaust.

PEMS 3, 7, 8 and MEL were operated by CE-CERT staff following manufacturer procedures for details on leak checks, zero checks and zero adjustments. PEMS 3 and 5 were operated by the manufacturer. Note that PEMS 8 was part of the MEL lab, calibrated against diesel exhaust and used regularly to indicate the expected concentration range. These values have helped to set dilution ratios and sampling times when measuring diesel sources of unknown PM concentration. However, it is more common for PEMS 8 to be calibrated with Arizona dust and used for ambient monitoring. For the purposes of this study, all figures and tables show the PEMS 8 values with the manufacturer calibration and uncorrected for the UCR correlation.

CE-CERT staff analyzed mass based PM emissions calculations using the FRM exhaust volume and PEMS mass concentration for PEMS 5, 7 and 8. Only PEMS 3 performed independent PM mass-based calculations. The FRM mass calculations used the same CVS volume and secondary dilution-ratio as the PEMS, but in addition required filter sample flows and filter mass.

		Mfg Max	Mfg Dynamic	Mfg Maximum		Mfg Size
ID	Output	Sample Rate	Range	Range	Mfg LDL	Range
<b>FRM</b>	Teflon_ug	$1/c$ ycle	100	1ma	10 <sub>ug</sub>	$< 2500$ nm
EC/OC	Quartz_ug	$1/c$ ycle	1000	< 10mg	10 <sub>uq</sub>	$< 2500$ nm
PEMS3	Teflon_ug	$1/c$ ycle	250	2.5 <sub>mg</sub>	10 <sub>uq</sub>	n/a
PEMS5	uq/m3	20Hz	1000000	10000 mg/m3	$\lt 1$ ug/m $3$	n/a
PEMS7	mq/m3	20Hz	>5000	$50$ mg/mg $3$	~5uq/m3	n/a
PEMS8	mg/m3	1Hz	100000	100mg/m3	$1$ ug/m $3$	100-2500 nm

**Table 5-1. Specifications for the Different PEMS Used During the Mobile Source Testing.** 





## **5.2 Results**

The PM was sampled concurrently with the gaseous emission tests and the PM instruments were only evaluated over a complete integrated cycle. Table 5-3 shows the complete test results for each of the PEMS and FRM. Additional details of the results are provided in Appendix K. The blank spaces are invalid test points removed due to things such as lost filter flow, air compressor failure, serial port buffer overruns, instrument voltage dropping below manufacturer recommended levels, PEMS not operating properly as identified by PEMS technician, and frozen serial data from serial communication errors. The quartz filters for EC mass were sampled on only a few tests, which accounts for most of the blank EC mass values.

After a review of the FRM data, the manufacturer of PEMS 5 was concerned that their data was not representative of instrument design. They concluded that the signal recorded by CE-CERT was in error compared to their duplicate copy of the same signal. PEMS 5 output a volt signal to the CE-CERT for ease of post processing and time alignment. It appears there was a problem with the output signal that made the CE-CERT PEMS 5 data to be in error. PEMS 5 post processed all their own mass integrated numbers and second by second numbers and submitted them back to CE-CERT. After a review and discussion between ARB, the manufacturer and CE-CERT, it was decided to replace the CE-CERT PEMS 5 data by data processed from PEMS 5 manufacture. All figures and tables show PEMS 5 data processed by PEMS 5 manufacture.

**Table 5-3. Emissions of Total and EC Mass for PEMS and FRM Tested Over Various Cycles.** 

	PM g/cycle							
Trace	FRM	PEMS <sub>5</sub>	PEMS <sub>7</sub>	PEMS <sub>8</sub>	PEMS <sub>3</sub>	EC		
<b>UDDS</b>	3.55	5.65	3.09	5.16	3.14	2.45		
<b>UDDS</b>	3.61	6.39	3.08	5.26	3.03	2.55		
<b>UDDS</b>	3.77	6.77	3.26	5.54	3.24			
<b>UDDS</b>	3.81	5.86	3.34	5.41	3.24			
<b>UDDS</b>	4.00	6.10	3.48	5.73	3.43			
<b>UDDS</b>	4.07	5.60		5.16	3.02			
<b>UDDS</b>	3.87	5.34		4.99	2.87			
50CRUISE	4.30	6.74	3.53	5.64	4.14	3.08		
50CRUISE	4.20	6.45	3.46	5.55	3.82	3.03		
50CRUISE	3.98	6.31	3.33	5.32	3.95	2.86		
50CRUISE	4.16	6.29	3.38	5.51	3.76	2.84		
50CRUISE	4.29	6.30	3.44	5.60	3.87	2.97		
50CRUISE	4.95	6.10		5.40	3.66	2.74		
50CRUISE	3.42	6.74	3.51	5.60	3.80	2.42		
NTE 1290	4.12	6.02			4.18			
NTE 1290	3.44	5.52	2.90	4.88	3.25	2.42		
NTE_1500	4.49	6.43			6.26			
<b>NTE 1500</b>				6.52	4.44			
NTE 1770	7.81	8.25		9.69	11.67			
<b>NTE 1770</b>		10.15		10.53	8.28			
NTE_Stepped				6.62	4.76			
NTE_Stepped	5.58	6.23	4.20	7.70				
NTE_Stepped	5.38	6.00	4.18	7.41				
NTE_Stepped	5.75	7.42	4.40	8.33	4.84	3.61		
NTE_Stepped	6.59	8.65	5.28	10.42	5.15			
NTE_Stepped	7.52	7.11	6.01	10.35	6.44			

The PM mass emission results are presented in Figure 5-1, with error bars representing 95% confidence intervals for multiple measurements. In comparison with the FRM filter measurements, PEMS 3 and 7 agree within 15% to 25%; and PEMS 5 and 8 were within 10% to 70%.

It should be noted that the NTE cycles generally had a larger standard deviation than the UDDS and 50 mph cruise. The larger standard deviation is partly due to the NTE cycles being performed twice as compared with the seven repeats for the time-velocity traces. The NTE cycles were also not designed as strict speed versus time traces so there can be some variability in how the driver approaches the NTE events with different test iterations. The NTE cycles are included in the analysis since each point has significance for correlation comparisons.



**Figure 5-1. Mass Emissions for PEMS 3, 5, 7, 8 and FRM.** 

The correlation plot between the FRM and the PEMS is provided in Figure 5-2. Looking at the plot, the linear regression for PEMS 3 was closest to the parity line, but showed a correlation coefficient of 0.68. PEMS 3 response was typically 5% to 20% lower then the FRM except for two points that were around 40% higher then the FRM. PEMS 7 results were consistently below the FRM, but showed a high correlation coefficient of 0.95. PEMS 8 also had a high correlation coefficient of 0.9, but was consistently higher then the FRM. The good correlation for PEMS 7 and 8 indicate that the agreement for these instruments might be improved if calibrated against the gravimetric filter weights. PEMS 5 regression coefficient was greater than the FRM and had the lowest correlation of 0.53 compared to the FRM.

There are some sources of systematic error or bias that should be considered when comparing the PEMS and the FRM. Of the PEMS only 8 and the FRM sampled through an inlet classifier prescribed by the 2007 standards. The inlet classifier would remove a small fraction of the total mass ( $\sim$  5%) attributed to the particles  $\geq$ 2.5µ. As a consequence PEMS 3, 5, and 7 would be biased slightly higher when compared with 8 and the FRM. Another bias results from the filter face temperature. The FRM measures PM with filters at 47 °C where all the PEMS inlets are maintained at near ambient temperatures of 25 °C. Previous studies have shown the PM mass is increased by approximately 11% when the filter face temperature is reduced from  $47^{\circ}$ C to  $25^{\circ}$ C. This effect would bias the FRM low as compared to the PEMS. The classifier and temperature biases should have both lowered the FRM mass compared to the PEMS. Only PEMS 8 was consistently higher than the FRM, but the difference was considerably greater than the 15% bias that could have been attributed to the combined classifier and temperature effects. The work suggests that other sources of error need to be investigated.



**Figure 5-2. Plot of FRM PM Mass Correlation with PEMS PM Mass.** 

Statistical comparisons of the test cycles results are provided in Appendix L. The statistical analyses included a paired and unpaired t-test and an F-test for comparison of experimental variance. The statistical comparisons for the cycles showed that in all cases the differences between the mass emission rates for the different instruments were statistically significant at greater than a 95% confidence level. The statistical significance of the F-test results differed depending on the specific test iteration and the specific emissions component.

The measurement principles incorporated into the design of PEMS 5, 7, and 8 are such that the instruments are expected to better predict the PM when the EC is a high percentage of the PM mass. The other techniques, including the FRM, are designed to measure the total PM. Based on the data in Table 5-3, the EC represented approximately 80% of the total carbon (TC) mass and did not vary significantly for the different cycles.

For this calculation, the TC mass was determined by the sum of EC and 1.2\*OC, the later factor added to account for oxygen, hydrogen and other elements related to the carbonaceous material. The high EC/TC ratio was as expected based on the previous studies that showed a high conversion over a diesel oxidation catalyst such as included on the exhaust system of the test vehicle.

Figure 5-3 shows the EC/TC ratios and a comparison of FRM total and EC masses to the total mass determined by PEMS 5, 7, and 8. From Figure 5-3, the mass measured by PEMS 7 shows a PM mass that was between that of the FRM gravimetric and FRM EC mass while PEMS 5 and 8 was always higher.



**Figure 5-3. PM Mass Emissions for PEMS 5, 7, 8 and FRM. FRM Includes EC mass. Values Represent the EC/TC Average Ratio for Each Cycle.** 

Figure 5-4 presents the same data as a correlation plot of the EC mass and the total mass measured by PEMS 5, 7, and 8. The correlation coefficient between the EC mass and PEMS 5, 7, and 8 mass was less than the correlation with gravimetric FRM mass. The correlation between EC mass and the FRM gravimetric mass was 0.76. The low correlations could be due to smaller data cluster in the subset of tests where EC measurements were made.



**Figure 5-4. PEMS correlation to FRM by PM EC mass.** 

Figures 5-5 and 5-6 are plots of second-by-second data for one of the NTE steady state cycles and one of the UDDS cycles. These figures represent a significant advance over the current FRM and PEMS 3 that can collect only integrated data and miss the transitions in diesel operation when the air fuel ratio is varied considerably during the driving schedule. A good example of the highly transient nature of real world driving and PM emissions can be viewed during the transitions. Further note that the real-time data shows similarities between the instruments with respect to when peak PM emissions are formed as well as differences in the magnitudes of the PM instruments. Similar trends to the bar charts and correlation plots can be seen in comparison to PEMS 7 where PEMS 5 and 8 were higher. Notice that PEMS 5, 7 and 8 all show relatively the same transient spike with a change in engine torque. Also notice how transient the torque is on the UDDS large hill and all the PM oscillations during this transient torque condition.



**Figure 5-5. Primary Diluted PM Concentration for PEMS 5, 7, and 8 for the NTE 1290 RPM Cycle at 36% and 60% Rated Torque. Note: e-Torque is an ECM Broadcast Measurement.** 



**Figure 5-6. Primary Diluted PM Concentration for PEMS 5, 7 and 8 for the Middle Hill of the UDDS Cycle. Note: e-Speed and e-Torque are ECM Broadcast Measurements.** 

# **6.0 Discussion of Overall Results**

The goal of this project was to provide data on the performance of PEMS as alternatives to EPA reference methods for measuring real-world, on-board, in-use emissions from diesel engines. Comparative measurements were made for the concentrations of audit gases ('true values") and emissions from both a stationary source (backup generator) with true steady state engine operation and a mobile source (heavy-duty diesel truck) on a chassis dynamometer with quasi-steady-state and transient engine operation. These measurements were intended to provide quantitative insight concerning the use of PEMS for regulatory applications, such as the Not-To-Exceed (NTE) regulations and emission inventories.

A total of seven PEMS manufacturers were represented in this research program and as one manufacturer provided two instruments, a total of eight instruments were tested. Three instruments measured solely particulate matter (PM), three measured solely gaseous emissions and the last two units measured both PM and gaseous emissions. Several points need to be made. First, the highest priority was establishing the data for the gaseous instruments and second, not all the PEMS systems were used for both the stationary and the mobile source tests.



Emission Measurement Tests –Stationary Source/Backup Generator

The first phase measured emissions from a diesel backup generator and was intended to represent true steady-state engine operation. Flow rates and emission measurements were measured with four PEMS and compared with the values from the FRMs while the BUG operated at four load points. Results for  $NO<sub>x</sub>$  showed the deviation range depends strongly on the PEMS and the load. PEMS #2 and #3 show the best agreement with the FRM, ranging from about 1 to 10% of the FRM. Reasons for deviations observed with the other PEMS were discussed in the report. Deviations for THC and PM were quite high suggesting that more development is needed.

Emission Measurement Tests –Mobile Source/Heavy, Heavy-duty Diesel Truck

The second phase measured emissions from a heavy-duty tractor operated on ARB's chassis dynamometer and was intended to represent quasi steady-state and transient engine operation. Flow rates and emissions were measured from a 475 hp Caterpillar C-15 ACERT engine certified to the 2.5 g/bhp-hr  $NO<sub>x</sub> + NMHC$  and 0.1 g/bhp-hr PM standards with eight PEMS and compared with values from FRMs. Three of the PEMS measured only PM, three PEMS measured only gases and the remaining PEMS measured both gases and PM.

Testing during this phase was carried out following a carefully planned series of quasi steady state and transient driving schedules. These four driving schedules included:

- 1) Three-mode steady-state NTE test cycle run.
- 2) Stepped NTE test cycle
- 3) The Urban Dynamometer Driving Schedule (UDDS)
- 4) The 50-mph Cruise mode from the ARB 5-mode HHDDT test cycle

While data were collected for four driving schedules the discussion focuses on the emissions results for only the 3-Mode steady state NTE cycle as the results for other cycles were generally comparable. Results depended strongly on the PEMS. For PEMS2 the  $NO<sub>x</sub>$  emission rates were approximately 3 to 17% higher than those for the FRM in g/bhp-hr units relative to an NTE threshold of 2.0 g/bhp-hr. PEMS 4 showed larger differences compared to the FRM, with percent differences ranging from 26 to 47%.

The  $CO<sub>2</sub>$  emissions were used to determine fuel specific (fs) emissions and results indicated that during steady state NTE events,  $CO<sub>2</sub>$  emissions for PEMS2 were generally about 5% higher than those of the FRM and PEMS4 values were 9 to 30% higher than those the FRM. Further analysis is offered in the main report.

Emission Measurement Tests –Mobile Source & PM-capable PEMS

One of the most interesting results from the research was the exploratory survey of the commercial PM-capable PEMS applied to a mobile source following transient driving schedules. The strong point of the results was the clarity which the PEMS can follow transitions in a diesel driving schedule and the magnitude of the PM releases while the engine dramatically changes the air/fuel ratio. However, these results must be mitigated with the observation that more work is needed to better understand the mismatch between the total mass measured by the FRM and PEMS when integrated over a number of cycles.

Next Steps

The original scope for this project included four main emissions testing tasks: 1) backup generator testing of all participating PEMS, 2) heavy-duty truck chassis dynamometer testing of all participating PEMS, 3) On-vehicle, over-the-road testing of all participating PEMS using pre-selected test routes, and 4) on-board, over-the-road testing of "the most suitable" PEMS under conditions meant to mimic the actual in-use compliance program (i.e., real vehicles in actual revenue service). However, since this project began, the ARB, US EPA and heavy-duty engine companies (as represented by the Engine Manufacturers Association) agreed to a measurement allowance program to determine exactly what numerical values should be assigned, as referenced in the Memorandum of Agreement (MoA) signed by the above parties in June, 2005. The original scope of this project is being revised so that the resources originally allocated for this project can reallocated to support this measurement allowance program. The measurement allowance program will be examining topics and issues similar to those studied in this project - how PEMS compare against the reference methods, but will do so in a more focused and extensive manner. Specifically, all factors expected or suspected to influence PEMS emissions measurements such as vibration, ambient conditions (such as changes and variations in pressure, temperature. and humidity, ambient background HC concentrations), radio frequency and electromagnetic interference, etc. will be studied with the goal of developing actual, pollutant-specific, numerical measurement allowances for  $NO<sub>x</sub>$ , THC, and CO. In addition, the measurement allowance program will also quantify variability between engine dynamometer measurements of engine torque and speed. The current project scope is being revised to support this measurement allowance program. Specifically, CE-CERT will: 1) perform a Code of Federal Regulations Part 1065 audit of the CE-CERT mobile emissions laboratory, 2) perform side-by-side testing at SwRI, and 3) perform a validation of the Monte Carlo model used to develop the measurement allowance by collecting over-the-road emissions data using the MEL.

# **7.0 Conclusions and Recommendations**

This project was borne out a need to identify the accuracy and precision of PEMS as such systems were being used to accumulate in-use data on emissions from buses, trucks and locomotives and other equipment for inventory purposes. Furthermore, new regulations requiring on-board, in-use measurement and compliance with the NTE rules were being issued and these parties also needed the information on accuracy and precision. Finally there is the realization that the measurement from large fixed laboratories and PEMS are both needed to advance and improve data used for preparing SIPs. Thus a broad stakeholder group was easily formed to serve as advisers for the research on a project destined to help improve the understanding of the comparative performance of traditional FRMs with PEMS.

An important perspective when viewing these findings is that PEMS are evolving products that use a wide range of different measurement principles and operating practices. Thus as expected, different PEMS show varying degrees of agreement with standard products and practices specified for FRMs; for example, a  $NO<sub>x</sub>$  analyzer, whose specifications and operation are outlines in great detail in the CFR. Further, the results from this program represent the best equipment and best operator. Thus it is best to state that results from this research represent a snapshot in time since the technology and capability of PEMS will continue to evolve as the product moves towards standardization.

Given these caveats, a number of conclusions can be reached based on the results of this study:

- Measured flow rates were in good agreement with the FRM, except for one PEMS unit with a failed part. One of the more creative approaches PEMS#1 who calculated the flow rate based on engine operation and gas law principles. Their results best matched the FRM for the stationary source.
- For  $NO<sub>x</sub>$  emissions, the best performing PEMS unit was within 5-10% of the FRM. This represented 5-25% of the 2.0 g/bhp-hr  $NO_x$  standard depending on the specific test and operating condition. The other PEMS showed larger differences for  $NO<sub>x</sub>$  with some showing significantly larger deviations from FRM.
- For  $CO<sub>2</sub>$  emissions, the best performing PEMS were able to measure within 5% of the FRM. In some cases, considerably larger deviations were found in  $CO<sub>2</sub>$  when there were errors in the measurement of exhaust flow rate, calibration or software.
- On an absolute basis, THC emissions showed considerably larger deviation than other emissions components due to the low concentrations of THC emissions. From an NTE perspective, the deviations between the PEMS and FRM were less than 5% of the NTE threshold standard for THC for the chassis dynamometer testing, with larger deviations seen for the stationary source tests. Results suggest that additional development is needed for accurate in-use THC emission values.
- The deviations between the PEMS and FRM varied for CO. One PEMS was within 5% of the FRM during the stationary source testing. For most other test scenarios, the differences between the FRM and PEMS ranged from 10-50% or higher. From an NTE perspective, the deviations represented less than 2% of the NTE CO threshold for nearly all test conditions. Additional accuracy may be needed to obtain in-use emissions measurements for CO for use in emissions inventories.
- A wide range of exciting new technologies were investigated for measuring PMcapable PEMS. Several of the instruments based their design on elemental carbon. As such, not all instruments were directly comparable with the FRM PM mass. The best PEMS for measuring PM were within 15-25% of the PM mass emissions measured by the FRM, with considerably larger deviations seem for some other instruments. It is expected that further development of PM PEMS will be needed to meet the regulatory challenges of measuring the low PM levels that will be found for engines equipped with after treatment devices.
- Results suggest that more comprehensive QA/QC procedures for in-field operation are needed to ensure reliable operation from a range of PEMS instruments. In the present program, issues such as problems measuring calibration gases were identified, suggesting the need for improved procedures for in-field units.

## Recommendations

The results of this research show that of the four gaseous PEMS that PEMS2 compared better with the FRM for  $NO_x$  and  $CO_2$  than other units. An exploratory trial of PMcapable PEMS showed great promise for real-time PM measurements but more development is needed. The following recommendations are made based on the observations in this research:

- More attention is needed to establish detailed use protocols and performance limits as specified for a CFR method. For example, the CFR specifies a particular type of analyzer for  $NO<sub>x</sub>$  as well as the specifications.
- Even the best PEMS with an experienced operator had problems so PEMS require experienced operators.
- It is suggested that more frequent calibrations are needed to establish confidence intervals and ensure reliable operation of the PEMS during sampling in the field.
- For the best progress, measurement programs must be collaborative in design and analysis between the PEMS manufacturer, ARB, EPA, EMA and the contract lab making measurements.

Looking ahead, we expect PEMS confidence limits to improve from that observed in this research as a result of the ARB/EPA/EMA's Measurement Allowance Program.

# **Appendix A – Background Information on UCR's Mobile Emission Lab**

Extensive detail is provided in Reference 2; so this section is provided for those that may not have access to that reference. Basically the mobile emissions lab (MEL) consists of a number of operating systems that are typically found in a stationary lab. However the MEL lab is on wheels instead of concrete. A schematic of MEL and its major subsystems is shown in the figure below. Some description follows.



## **Major Systems within the Mobile Emission Lab**

The primary dilution system is configured as a full-flow constant volume sampling (CVS) system with a smooth approach orifice (SAO) venturi and dynamic flow controller. The SAO venturi has the advantage of no moving parts and repeatable accuracy at high throughput with low-pressure drop. As opposed to traditional dilution tunnels with a positive displacement pump or a critical flow orifice, the SAO system with dynamic flow control eliminates the need for a heat exchanger. Tunnel flow rate is adjustable from1000 to 4000 scfm with accuracy of 0.5% of full scale. It is capable of total exhaust capture for engines up to 600kW. Colorado Engineering Experiment Station Inc. initially calibrated the flow rate through both SAOs for the primary tunnel.

The mobile laboratory contains a suite of gas-phase analyzers on shock-mounted benches. The gas-phase analytical instruments measure  $NO<sub>x</sub>$ , methane (CH<sub>4</sub>), total hydrocarbons (THC), CO, and  $CO<sub>2</sub>$  at a frequency of 10 Hz and were selected based on optimum response time and on road stability. The 200-L Tedlar bags are used to collect tunnel and dilution air samples over a complete test cycle. A total of eight bags are

suspended in the MEL allowing four test cycles to be performed between analyses. Filling of the bags is automated with Lab View 7.0 software (National Instruments, Austin, TX). A summary of the analytical instrumentation used, their ranges, and principles of operation is provided in the table below. Each modal analyzer is timecorrected for tunnel, sample line, and analyzer delay time.



#### **Summary of gas-phase instrumentation in MEL**

Quality Assurance and Quality Control Requirements

Internal calibration and verification procedures are performed regularly in accordance with the CFR. A partial summary of routine calibrations performed by the MEL staff as part of the data quality assurance/quality control program is listed in the table below. The MEL uses precision gas blending to obtain required calibration gas concentrations. Calibration gas cylinders, certified to 1 %, are obtained from Scott-Marrin Inc. (Riverside, CA). By using precision blending, the number of calibration gas cylinders in the lab was reduced to 5 and cylinders need to be replaced less frequently. The gas divider contains a series of mass flow controllers that are calibrated regularly with a Bios Flow Calibrator (Butler, New Jersey) and produces the required calibration gas concentrations within the required  $\pm 1.5$  percent accuracy.

In addition to weekly propane recovery checks which yield  $>98\%$  recovery, CO<sub>2</sub> recovery checks are also performed. A calibrated mass of  $CO<sub>2</sub>$  is injected into the primary dilution tunnel and is measured downstream by the  $CO<sub>2</sub>$  analyzer. These tests also yield  $>98\%$ recovery. The results of each recovery check are all stored in an internal QA/QC graph that allows for the immediate identification of problems and/or sampling bias.

An example shown below is for propane mass injected into the exhaust transfer line while sampling from raw and dilute ports (three repeats) to evaluate exhaust flow measurement on steady state basis (duration  $= 60$  sec, Date completed January 2005).







# **Sample of Verification and Calibration Quality Control Activities**

# **Appendix B – Descriptions of PEMS Instruments**

## **PEMS1 – CATI Montana System**

**1.** The CATI OEM 2100 PEMS measures second-by-second mass emissions from vehicle tailpipes with electronically controlled sparked ignition and compression ignition engines. The unit provides HC, CO, CO2, NOx and O2 readings for gasoline-powered vehicles and NOx, CO, CO2, O2, PM readings for diesel vehicles.

The unit provides second by second emissions, fuel consumption, vehicle speed, engine rpm and temperature, throttle position, and other parameters.

The CATI PEMS includes: Touch-screen computer (256MB RAM, USB, Serial, Parallel, Network Ports) · Dual Gas Analyzer -HC (gasoline only), NOx, CO, CO2, O2 · Light-Duty Engine Scanner · Heavy-Duty Engine Scanner · Sensor Array (for nonelectronically controlled vehicles) · Particulate Matter (PM) Monitor (diesel only); Weatherproof Case · Keyboard · Back-up Battery.

The unit is weighs 44lbs. The system uses power directly from a vehicle's 12V or 24V electrical system, consuming 8 A at 12V DC, or can be powered by AC in the case of stationary testing.

Engine data can be sensed directly using an array of analytical sensors. This method involves attaching several analog sensors to the engine itself.

For vehicles with a supported computer diagnostic port, engine and vehicle data is acquired using this interface. The unit is equipped with ECU scanners that will communicate with the ECU and obtain any needed engine parameters. The two most common engine scanners that can be incorporated are the heavy duty scanner and a light duty scanner.

The diagnostic port interface cable is routed directly to the unit from the port connector. For sensor array installations, sensors are installed on the applicable engine systems and are then routed to the unit.

## **2.** Instrument Range

The concentrations of HC,  $CO$ ,  $CO_2$ ,  $O_2$ ,  $NO_x$  and PM in the exhaust gas are determined by following methods and have the following ranges:

• Non-Dispersive Infrared (NDIR) analyzer for hydrocarbon (HC), carbon monoxide (CO) and carbon dioxide (CO<sub>2</sub>) measurement of  $0 - 2000$  ppm hexane  $(C_6)$  with an accuracy of  $\pm 4$  ppm abs. or  $\pm 3$  % rel., 0 – 10.00 % CO with an accuracy of  $\pm 0.02$  % abs. or  $\pm 3$  % rel., and  $0 - 16.00$  % CO<sub>2</sub> with an accuracy of  $\pm 0.3$  % abs. or  $\pm 3$  % rel.

- Electrochemical sensors for oxides of nitrogen  $(NO_x)$  and oxygen  $(O_2)$ measurement of  $0 - 4000$  ppm NO with an accuracy of  $\pm 25$  ppm abs. or  $\pm 4$  % rel. and 0- 25.00 %  $O_2$  with an accuracy of  $\pm 0.1$  % abs. or  $\pm 3$  % rel.
- Particulate matter (PM) concentration is measured using laser light scattering.

The PM range is from ambient levels to low double digit opacity levels. The accuracy of PM measurement has not be quantified at this point. Response is dependent on size and composition of the particles. As a result, there may be a systematic measurement bias associated with each individual vehicle. This bias is minimized by the use of virtual impactors which eliminate course mode particles (1-2 microns). The sampling system allows for small nuclei-mode particles (< 100 mm) to conglomerate or condense onto soot particles. This narrows the particle size distribution, allowing for a more accurate conversion of the reading to particulate mass concentration and PM mass emissions

**3.** The CATI PEMS measures ppm/second emission data. Theoretical exhaust flow is calculated using engine parameters read from the vehicle's engine control unit or the sensor array. Emission results are calculated by combining the theoretical exhaust flow and the collected ppm/second emission data. Results are reported in grams/second format.

From the intake air mass flow, known composition of intake air, measured composition of exhaust, and user supplied composition of fuel, a second by second exhaust mass flow is calculated. This calculation is proprietary, but involves a mass balance equation, whereby matter coming into the engine must equal that matter coming out of the engine. Multiplying the exhaust mass flow by the concentrations of different pollutants yields grams per second emission data.

Engine power output can be estimated based on ECU torque readings and/or using the fuel consumption and engine rpm data, and the manufacturer's *brake specific* fuel consumption charts. All measured pollutants can then be calculated on a *brake specific*  basis. This information can be used for NTE analysis.

#### **PEMS2 – Semtech D**

SEMTECH-D is a complete, fully integrated portable emissions measurement system (PEMS) for testing all classes of diesel-powered vehicles and equipment under real-world operating conditions. SEMTECH-D measures diesel emissions at the tailpipe, engine-out, or at any stage of after-treatment. A data logger records the vehicle emissions, environmental conditions, and the output of a vehicle's on-board electronic control system to compact flash removable storage while the vehicle is in operation. The optional exhaust mass flowmeter and GPS are also fully integrated with the SEMTECH-D data logger and post-processing software. Engine and vehiclerelated parameters are combined with gaseous emissions on a real-time basis to determine in-use emissions levels in  $g/sec$ ,  $g/g$ -fuel,  $g/Bhp-hr$ , and  $g/mile$ . Not to Exceed (NTE) vehicle operation and emissions results are also determined on a real-time basis. Test results can also be viewed subsequently with the user-configurable post-processor application.

Access to the central processor is provided through LabView™ PC host software. The user interface is designed to provide immediate feedback to the user. There are over 150 different fault codes that the SEMTECH will automatically report to the user if a problem occurs. In addition, there are 24 warning codes that will also automatically be reported when potential problems exist. They indicate to the user when to change filters, when to change the FID fuel bottle, when to zero the instrument. In addition, many of the routine tasks that are required to operate the system are fully automated, requiring minimal effort for the user.



The SEMTECH-D system comprises of eight individual analyzers, all integrated into a single package and controlled from a central processor/data logger. The following table describes the subcomponents and system features.



#### **PEMS3 – RAVEM System**

The Ride-Along Vehicle Emissions Measurement (RAVEM) is capable of measuring PM as well as NOx, CO, and  $CO<sub>2</sub>$ . Optional capabilities also allow the measurement and quantification of total hydrocarbons (THC), sulfur dioxide  $(SO<sub>2</sub>)$ , ammonia (NH<sub>3</sub>), and nitrous oxide  $(N_2O)$ , as well as individual species of volatile organic compounds (VOC) and carbonyls such as formaldehyde, acetaldehyde, and acrolein.

Principals of Operation

The RAVEM system is described in two published papers (1,2). As Reference **Error! Bookmark not defined.** explains in more detail, the RAVEM system is based on proportional *partial-flow* constant volume sampling (CVS) from the vehicle exhaust pipe. The CVS principle is widely used for vehicle emission measurements because the air dilution and total flow arrangements are such that the pollutant *concentration* in the CVS dilution tunnel is proportional to the pollutant *mass flow rate* in the vehicle exhaust.

The total pollutant mass emissions over a given driving cycle are equal to the integral of the pollutant mass flow rate over that cycle. In a CVS system, this integrated value can readily be determined by integrating the concentration measurement alone. The CVS flow rate enters into the calculation as a constant multiplier. The integration of pollutant concentration can be accomplished either numerically or physically. For gases, the RAVEM system uses both numerical and physical integration. Concentrations of NOx,  $CO<sub>2</sub>$ , and  $CO$  in the dilute exhaust gas are measured and recorded second-by-second during each test. In addition, integrated samples of the dilute exhaust mixture and dilution air are collected in Tedlar® bags during the test, and analyzed afterward for NOx,  $CO<sub>2</sub>$ ,  $CO$  and (optionally) other pollutants.

In CVS sampling for particulate matter, sample integration is accomplished physically - by passing dilute exhaust mixture through a pre-weighed filter at a constant, controlled flow rate. The weight gain by the filter is then divided by the volume of mixture passed through it to yield the average particulate concentration over the test cycle.

A schematic diagram of the RAVEM system is shown below. Except for the isokinetic sampling system at the top of the figure, this diagram closely resembles a conventional single-dilution CVS emission measurement system.

Conventional emission laboratory methods defined by the U.S. EPA (4) and California ARB (5) utilize full-flow CVS, in which the entire exhaust flow is extracted and diluted with air. However, the large amounts of dilution air required make full-flow CVS impractical for portable systems.

The principle of the RAVEM sampling system is as follows: the RAVEM's sampling system extracts and dilutes only a small, constant fraction of the total exhaust flow. The dilution air requirements and dilution tunnel size can thus be reduced to levels compatible with portable operation. The patented isokinetic proportional sampling system<sup>5</sup> continuously adjusts the sample flow rate so that the flow velocity in the sample probe is equal to that of the surrounding exhaust. Since the velocities are equal ("isokinetic"), the ratio of the flow rates in the exhaust pipe and the sample probe is equal to the ratio of their cross-sectional areas.

Pollutant concentration measurements in the RAVEM system follow the methods specified by the U.S. EPA (US CFR Vol 40 Part 86) and ISO standard 8178. The pollutants measured are:

- Oxides of Nitrogen (NOx) by chemilumenescent analysis of the dilute exhaust sample. The zero-100 ppm range is normally used, but ranges from 0-10 to 0-3000 ppm are available;
- Carbon monoxide (CO) and carbon dioxide  $(CO_2)$  by non-dispersive infrared analysis of the dehumidified dilute exhaust sample. The 0-200 ppm range is normally used for CO, but a 0-500 ppm range



is available. For  $CO<sub>2</sub>$ , the 0-6000 ppm range is normally used, 0-2000 and 0-10,000 ppm ranges are also available;

• Particulate matter (PM) is measured by passing the dilute exhaust sample through pre-weighed 47 mm filters of Teflon-coated borosilicate glass fiber, followed by post-conditioning and reweighing. The minimum detectable PM filter mass is approximately 10 micrograms, the maximum practical mass on the filter is more than 3000 micrograms. Filter and CVS flow rates can also be adjusted to increase PM sensitivity or avoid PM overloading.

References

<sup>1</sup> C.S. Weaver and L.E. Petty "Reproducibility and Accuracy of On-Board Emission Measurements Using the RAVEM™ System ", SAE Paper No. 2004-01-0965, March, 2004.

<sup>2</sup> Weaver, C.S. and M.V. Balam-Almanza, "Development of the 'RAVEM' Ride-Along Vehicle Emission Measurement System for Gaseous and Particulate Emissions", SAE Paper No. 2001-01-3644.

<sup>3</sup> 40 CFR 86, Subpart N "Emission Regulations for New Otto-Cycle and Diesel Heavy-Duty Engines; Gaseous and Particulate Exhaust Test Procedures"

<sup>4</sup> "California Exhaust Emission Standards and Test Procedures for 1985 and Subsequent Model Heavy-Duty Diesel Engines and Vehicles" as amended on February 26, 1999, California Air Resources Board

<sup>5</sup> U.S. Patent No. 6,062,092. "System for Extracting Samples from a Stream", May 16, 2000.

#### **PEMS4 – Horiba OBS-2200**

The HORIBA OBS-2200 is an on-board emission measurement system capable of analyzing vehicle emissions during in-use operation. The OBS 2200 consists of vibration-proof gas analyzers, a laptop PC with software for system controlling and data logging, accessory sensors, and a Pitot tube tailpipe flowmeter. All gases are analyzed wet, without drying.  $CO$ ,  $CO<sub>2</sub>$  and water vapor concentrations are measured by a heated NDIR analyzer. The water measurement compensates for water vapor interferences. THC concentrations are measured by a heated FID analyzer (190ºC), and NOx concentrations are similarly measured by a heated CLD analyzer. The PC data logging software saves analyzer data, OBD ECU data, GPS data, tailpipe flow, and other external signals, as well as outputs from accessory sensors. The software provides time-trend profiles and integrated values for both mass emissions and fuel consumption. Software identifies NTE events and reports results.

#### Software Calculations:

The following data can be displayed in real time and logged to file:

- Concentration of CO,  $CO<sub>2</sub>$ , THC, and NOx;
- Exh. flow rate  $[m^3/min]$ ; Exh. temp.  $[°C]$ ; Exh. press.  $[kPa]$
- Amb. temp. [ºC]; Atm. press. [kPa]; Amb. humidity (relative humidity) [%]
- GPS velocity [km/h]; Altitude [m]; Position (latitude/ longitude)
- External inputs (optional)
- OBD inputs (optional)

The following items can be calculated and displayed in real time and logged to file using input data:

- Mass emission of CO,  $CO<sub>2</sub>$ , THC, and NOx [g/s, g/h]
- Fuel consumption [g/s]; Fuel economy [km/L, mile/L, L/100km, g/kWh or g/bhph] A/F (calculated by carbon balance method)
- 
- Power [kW] (calculated from engine speed and torque/%torque)

Setting values to be used in calculation are as follows:

- Time alignment delay of  $CO$ ,  $CO<sub>2</sub>$ , THC, NOx analyzer response,
- (Time aligned with exhaust flow rate)
- H/C, O/C and density of fuel
- H/C ratio of hydrocarbon in exhaust emission

The following Calculated Data (as integrated values) can be calculated, displayed, and logged to file using input data:

- Mass emission of  $CO$ ,  $CO<sub>2</sub>$ , THC and NOx [g]
- Fuel consumption [g]; Fuel economy [km/L, g/mile, L/100km, g/kWh or g/bhph]
- Running distance [km]
- Work [kWh]

A Maximum of 8 items from the following list can be displayed in real time in chart/graphical format:

- Concentration of  $CO$ ,  $CO<sub>2</sub>$ , THC, and NOx
- Exh. flow rate [L/min]; Exh. temp. [ºC]; Exh. press. [kPa]
- Amb. temp. [ºC]; Atm. press. [kPa]; Amb. humidity (relative humidity)[%]
- GPS velocity [km/h]
- AFR (calculated value)
- External inputs (optional)

# - OBD inputs (optional)

# **OBS-2200 EMISSIONS MEASUREMENT SYSTEM**



## OBS 2200 SYSTEM CONFIGURATION



#### **PEMS5 The Artium Technologies, Inc.** *LII 200*

The Artium Technologies, Inc. *LII 200* instrument used in these investigations consisted of a self-contained rugged optics enclosure which includes the laser and all components needed for operation of the instrument. The optical system consists of a computercontrolled automated laser beam energy generation, detection and adjustment system that maintains the laser light fluence at the sampling volume at optimum conditions. Optics are provided to transform the laser beam intensity into a near top-hat light intensity profile which facilitates uniform heating of the soot in the probe volume. Laser-induced incandescence (LII) employs a high-energy pulsed laser beam to rapidly (10 to 20 nanoseconds) heat the soot particles from the local ambient temperature to below 4000 K as they flow through the sample cell. The laser heating is independent of particle size, and the emitted light is nominally volumetric. Thermal emission, i.e., incandescence, from the particles is then recorded, using collection optics and photodetectors, as the particles slowly (1 to 2 microseconds) cool to the ambient temperature. Using appropriate calibration and analysis of the incandescence signal, the particle volume fraction and active surface area/primary particle diameter are estimated, where the former is obtained from the amplitude of the signal and the latter from the signals' rate of decay. The laser is typically pulsed at 20 Hz and a measurement is obtained for each laser pulse.

Prior to making measurements, calibration measurements are made using a known traceable radiance source at a known temperature, which provides calibration factors to relate the measured signals to absolute intensities. Auto-compensating LII (AC-LII), which is incorporated into the Artium LII 200, is based upon measurements of the particulate surface temperature determined by optical pyrometry. Incandescence at two or more independent wavelengths,  $\lambda$ , is recorded, and an average soot particle surface temperature in the probe volume is calculated by using the ratio of the observed signals (corrected for detection sensitivity) and the known soot particle absorption cross sections. This has the added advantage of permitting the laser light heating of the soot to be kept below the sublimation (vaporization) temperature so no significant soot volume is lost as a result of the measurement.

Soot (elemental carbon based particles emitted from combustion sources) absorbs and emits light predominantly on the scale of the primary particles. LII measurements are insensitive to liquid particles, because they absorb a negligible amount of laser energy compared to carbon and do not produce any incandescence signal. For carbon particles coated with volatile material, the latter will vaporize very early in the LII laser-heating period and AC-LII automatically compensates for the energy lost to evaporating the volatile fraction by measuring the particle temperature in real time, such that the measure of the soot particle is not affected. In general, it is reasonable to state that LII measures the volume fraction of solid carbonaceous material in the exhaust. Other constituents may include trace metals and ash. Although trace metals may contribute to the incandescence if they survive the peak temperatures, their concentration is typically so low that the contribution will be negligible.

The LII instrument measures the soot volume fraction (SVF) that is a volume concentration. The additional measurement of the exhaust flow rate leads to the timeintegrated emissions results.

## **Measurement Range:**

Using a density for soot of 1.9  $g/cm<sup>3</sup>$ , we have demonstrated a mass concentration minimum detection level of 10  $\mu$ g/std m<sup>3</sup> and a maximum mass concentration level of 10 g/std m<sup>3</sup>. This system can be extended to reach a minimum detection level of 1  $\mu$ g/std m<sup>3</sup>. These results can be obtained at the exhaust exit since dilution and conditioning of the exhaust are not needed when using LII. In these investigations the LII sample was drawn from the CVS dilution tunnel.

Transient measurements for a range of 2000:1 have been demonstrated to date. Using the LII 200 dynamic PMT (photomultiplier tube detectors) gain control, the instrument can reach 10,000:1. Larger measurement loading ranges requires a change of optical filters in the system which is done automatically when the software logic senses that it is required. This typically requires 0.5 to 2 seconds. It does not require operator intervention to do this. The instrument range is 10,000,000:1 but this cannot be achieved for a single transient that is shorter than the 0.5 to 2 seconds time specified because of the need to change the filters. A range of 2000:1 in concentration can be achieved on a measurement to measurement basis at 20 Hz.

Primary particle diameters of 5 nm to 100 nm have been measured with the LII 200, and do not represent the limits of the instrument.

## **Emissions Measurements:**

The cycle integrated soot mass emissions were calculated as follows:

- 1. The 20 Hz LII soot volume fraction (SVF) data was multiplied by the density of soot  $(1.9 \text{ g/cm}^3)$ , to provide a 20 Hz mass concentration of soot in the exhaust.
- 2. A 20 point average of the 20 Hz mass concentration data was calculated, to provide a 1 Hz mass concentration data rate.
- 3. The 1 Hz mass concentration data was multiplied by the CVS flow rate on a second by second basis to determine the total mass flux of soot particulates from the engine, again on a 1 Hz basis.
- 4. The 1 Hz mass flux of soot emitted by the engine was integrated over the relevant portion of driving cycle (relevant times provided by CE-CERT) to produce an integrated mass of soot particulates over the test cycle.



Schematic of the LII 200 Optical system



## **PEMS7 – AVL 483 Micro Soot Sensor**



# PRODUCT INFORMATION **AVI**



#### AVL 483 Micro Soot Sensor

Specifications:


#### **Calculating the emission when a diluter with a known, constant or slightly varying dilution ratio is used.**

For these calculations, *the time-resolved exhaust gas mass flow q<sub>mew</sub> in [kg/h] must be known.(1)* 

$$
\mathbf{M}_{\text{PM\_test}} = \sum_{i=1}^{n} \frac{c_i \cdot r_{d,i} \cdot q_{\text{mew,i}} \cdot \Delta t_i}{3600 \cdot \rho_0}
$$

 $ρ<sub>0</sub>$  being the density of the exhaust gas under standard conditions (0 °C, 1013 mbar).

It can be equated to the density of air,  $p_0$  (air) =1.293 kg/m<sup>3</sup>, within approx. 1 %.

If the dilution factor  $r_d$  = const., then with  $\Delta t$  =  $T/n$  = const the result is (2):

$$
\mathbf{M}_{P\mathbf{M}_{\text{atest}}} = \frac{\mathbf{T} \bullet \mathbf{r}_{\text{d}}}{3600 \bullet \mathbf{n} \bullet \rho_{0}} \sum_{i=1}^{n} c_{i} \bullet q_{\text{new},i}
$$

If If the dilution factor  $r_d$  is not constant, then the result is  $(3a)$ :

$$
M_{\rm PM\_test} = \frac{T}{3600 \cdot n \cdot \rho_0} \sum_{i=1}^{n} c_i \cdot r_{d,i} \cdot q_{\rm mew,i}
$$

Or, if a dilution-corrected value of the soot concentration is available ( as in AVL 483 plus diluter), then the result is (3b):

$$
\mathbf{M}_{P\mathbf{M}\_\text{test}} = \frac{T}{3600 \bullet \mathbf{n} \bullet \rho_0} \sum_{i=1}^{n} c_{i \cdot dil-corr} \bullet q_{\text{new},i}
$$

The sum has to be calculated point by point online or offline. Separating the sum

$$
\sum_{i=1}^n c_i,_{dil-corr} \bullet \sum \ q_{mew,i}
$$

into : **is not admissible**.

## **Exposure Monitoring**

# **DustTRAK'" Aerosol Monitor**

The DusrTRAK™ Aerosol Monitor measures aerosols in a wide variety of environments, from offices and industrial workplaces to outdoor environmental and construction sites. TSI's DUSTTRAK provides reliable exposure assessment by measuring partide concentrations corresponding to PMIO, PM2.5, PMl.0 or respirable size fractions.

The DUSTTRAK is a portable, battery-operated laser photometer which gives you a real-time digital readout with the added benefits of a built-in data logger. Suitable for clean office settings as well as harsh industrial workplaces and outdoor applications, the DUSTTRAK detects potential problems with airborne contaminants such as dust, smokes, fumes and mists.

The DUSTTRAK is easy to use, too. You can perform quick spot checks or you can program the advanced logging modes for long-term sampling. You can program the start/stop times, recording intervals and other parameters. You can even set up the instrument for continuous unattended operation.

The DUSTTRAK's new continuous analog output and adjustable alarm output allow remote access to real-time particle concentration data. Applicatioos indude site perimeter monitoring, ambient monitoring, process area monitoring and other remote uses. The alarm output with user-defined setpoint alerts you when upset or changing conditions occur. This feature allows you to program a switch closure at a concentration value of your choosing.



The DUSTTRAK provides a real-time measurement based on *90°* light scattering. A pump draws the sample aerosol through an optics chamber where it is measured. A sheath air system isolates the aerosol in the chamber to keep the optics clean for improved reliability and low maintenance.



#### **Specifications Model 8520 DustTRAK Aerosol Monitor**

Sensor Type Range

Resolution

Zero Stability

Particle Size Range Flow Rate Teraperature Coefficient

Operating Temperature Storage Temperature Operating Huraidity Time Constant Data Logging

Logging Interval Physical

External Dimensions

Instrument Weight Serial Interface

Power  $\rm AC$ Battery Battery Run-time

Analog Output Specifications<br>Analog Output Voltage 0<br>Analog Output Scaling<sup>1</sup> 0

Output Impedance Maximura Output Current



The DUSTTRAK comes complete with TSI's TRAKPRO<sup>TM</sup> Data Analysis Software to allow you to perform a more comprehensive analysis of your measurement results. This exclusive Windows<sup>®</sup>-based program helps you generate the detailed graphs and reports needed to effectively communicate your findings.

90° light scattering<br>0.001 to 100 mg/m<sup>3</sup> (Calibrated to ISO

±0.1% of reading or ±0.001 mg/m<sup>3</sup>,

whichever is greater<br>±0.001 mg/m<sup>3</sup> over 24 hours using

0.1 to approximately 10 micrometers

Adjustable 1.4 to 2.4 l/min (1.7 nominally)

+0.001 mg/m<sup>3</sup> per  ${}^{\circ}\text{C}$  (for variations from

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32° F to 120° F (0°C to 50°C)<br>-4° F to 140° F (-20°C to 60°C)<br>0 to 95% th (non-condensing)

Adjustable from 1 to 60 seconds

once/minute)

RS-232 1200 baud

Alkaline 16 hours

 $0 \text{ to } 5 \text{ VDC}$  $0$  to  $100$  mg/m<sup>3</sup> 0 to 10.0 mg/m<sup>3</sup>  $0$  to  $1.00 \text{ mg/m}^3$  $0 \text{ to } 0.100 \text{ mg/m}^3$ 

 $0.01$  ohtn

 $15 \text{ mA}$ 

 $8.7$  in.  $\times$  5.9 in.  $\times$  3.4 in.  $(221 \text{ mm} \times 150 \text{ mm} \times 87 \text{ mm})$ 

31,000 data points (21 days of logging

Adjustable from 1 second to 1 hour

3.3 pounds with batteries (1.5 kg)

AC adapter (included)<br>Four C-size alkaline batteries (included)

12103-1, A1 test dust)

10-second time-constant

ecifications are subject to change without notice. Windows is a registered trademark of the Microsoft Corporation



#### **TSI** Incorporated

500 Cardigan Road, Shoreview, MN 55126 USA<br>Tel: 651 490 2811 Toll Free: 1 800 874 2811 Fax: 651 490 3824 E-mail: answers@tsi.com TSI Germany-Tel: +49-241-523030 Fax: +49-241-5230349 E-mail: tsigmbh@tsi.com TSI Sweden-Tel: +46-18-52-70-00 Fax: +46-18-52-70-70 E-mail: tsi@tsi.se

P/N 2980077 Rev. C

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Alarra Output Specifications

Type Setpoint Range<sup>1</sup> Maximum Voltage Maximum Current Deadband Connector

Non-latching, MOSFET solid state (polarized)<sup>2</sup> analog switch<br>0.010 to 100 mg/m<sup>3</sup> 15 VDC 1 Amp<br>5% of alarm setpoint 4-Pin, Mini-DIN connector

 $\frac{1 \text{ User schedule through TRACPE}^{\infty} \text{Data Analysis Software.}}{2 \text{ See TSI Application Note ITI - 074 for important writing information.}}$ 

#### **Ordering Information**

#### Model Description

8520 The DUSTTRAK Aerosol Monitor and accessories includes: The LUST FAR Arenosol Monitor and accessoris inculated Batteries, TaRAPRO<sup>36</sup> Data Analysis Software, Filter, Computer Cable, 25-pin to 9-pin Adapter, Operation Service Manual, Calibration Certificate, 10 nm Nylon Dorr-Oli Miscellaneous Service Tools and Two-Year Warranty.

#### **Ontional Accessories**

Model Description 8520-1 Environmental Enclosure



**Korowent information** 



# **Appendix C – PEMS & FRM Measurements of Blended Audit Gases**



Table C-1. Blended Audit Gases for BUG Testing

C-2 Audit Bottle Results for the Chassis Dynamometer Testing

			<b>Concentration Measured</b>		Audit Bottle Standard (1% NIST)							
			Cal Type NOx ppm THC ppmC1 CO ppm CO2 % NOx ppm THC ppmC1 CO ppm CO2 %									
PEMS <sub>1</sub>	raw	n/a	n/a	n/a	n/a	662	83.85	203.3	5.06			
PEMS <sub>2</sub>	raw	670.8	77.90	230.00	4.980	662	83.85	203.3	5.06			
PEMS <sub>3</sub>	dilute	151.5	0.00	22.40	1.436	151.8	23.91	25.14	1.556			
PEMS <sub>4</sub>	raw	834.7	100.69	184.67	5.943	662	83.85	203.3	5.06			
<b>FRM</b>	raw	677.2	83.84	204.51	5.110	662	83.85	203.3	5.06			
<b>FRM</b>	dilute	150.5	23.99	24.58	1.558	151.8	23.91	25.14	1.556			

### **Appendix D – BUG Testing Individual Test Results for Each Measurement Device**

BUG Test Results for the FRM –Run 1 in grams per hour









#### -**BUG Test Results for the FRM –Run 2 in grams per hour**





								Mass Emission g/Hp/hr			Exh Flow <b>Ambient Conditions</b>				Concentration Measured (wet)			
<b>Test Name</b>	Mode	Time sec	Load Factor	Load kW	<b>THC</b>	CO	kNOx	NO <sub>2</sub>	CO <sub>2</sub>	<b>PM</b>	scfm		Temp C Baro mmHckH				THC ppm CO ppm NOx ppm CO2%	
200411030833	1	450	100%	365.7	0.1046	0.749	7.253		517.4	0.048	1026				50.36	186.6	1095.5	8.170
200411031011	1	450	100%	362.4	0.1155	0.816	7.488		519.5	0.039	1009				55.80	205.0	1140.4	8.271
200411031157	$\mathbf{1}$	450	100%	361.7	0.1495	0.962	7.429		515.2	0.036	980				74.96	247.1	1158.2	8.397
200411031323	1	450	100%	360.1	0.1701	0.882	7.683		518.7	0.038	984				84.81	221.4	1183.7	8.355
200411040805	1	450	100%	364.1	0.1063	0.736	7.165		511.1	0.040	1033				50.46	179.9	1074.8	8.017
200411040944	1	450	100%	363.2	0.0943	0.755	7.366		504.4	0.036	1026				44.99	186.1	1106.7	7.922
200411041120	1	450	100%	363.2	0.1080	0.771	7.343		505.1	0.034	1022				51.63	193.1	1107.2	7.962
			ave	362.9	0.1212	0.810	7.389	#DIV/0!	513.1	0.039	1012	#DIV/0!	#DIV/0!	#	59.00	202.7	1123.8	8.156
			stdev	1.8	0.0277	0.084	0.168	#DIV/0!	6.3	0.005	21.4	#DIV/0!	#DIV/0!	#	14.89	24.1	38.3	0.193
			COV	0.5%	22.9%	10.3%	2.3%	#DIV/0!	1.2%	0.118	2.1%	#DIV/0!	#DIV/0!		25.2%	11.9%	3.4%	2.4%
200411030833	$\overline{2}$	450	65%	240.4	0.1937	0.710	8.503		515.2	0.039	786.1				80.04	151.1	1093.1	6.924
200411031011	$\overline{\mathbf{c}}$	450	65%	240.0	0.2080	0.815	8.652		516.0	0.038	774.1				86.92	172.8	1125.9	7.020
200411031157	$\overline{\mathbf{c}}$	450	65%	243.9	0.1994	0.877	8.589		508.5	0.038	763.7				85.63	194.8	1149.7	7.116
200411031323 200411040805	$\overline{2}$ $\overline{2}$	450 450	65% 65%	241.4 241.7	0.2426 0.1151	0.848 0.699	8.882 8.371		511.4 504.4	0.038 0.040	763.5 792.4				103.51 47.05	182.6 148.5	1173.8 1076.6	7.065 6.780
						0.741	8.515										1095.4	
200411040944 200411041120	$\overline{\mathbf{c}}$ $\overline{2}$	450 450	65% 65%	240.9 240.9	0.1244 0.0982	0.752	8.393		498.5 498.5	0.038 0.037	788.4 784.6				50.67 40.11	158.7 161.5	1087.1	6.705 6.751
				241.3	0.1688	0.777	8.558	#DIV/0!	507.5	0.038	779.0	#DIV/0!	#DIV/0!	#	70.56	167.1	1114.5	6.909
			ave stdev	1.3	0.0553	0.070	0.174	#DIV/0!	7.3	0.001	11.9	#DIV/0!	#DIV/0!	#	24.31	17.1	36.3	0.165
			COV	0.5%	32.8%	8.9%	2.0%	#DIV/0!	1.4%	0.021	1.5%	#DIV/0!	#DIV/0!		34.5%	10.2%	3.3%	2.4%
200411030833	3	450	25%	90.4	0.2820	0.618	8.793		627.4	0.044	500.8				68.43	78.9	662.6	4.943
200411031011	3	450	25%	90.7	0.3323	0.645	8.943		624.9	0.044	494.3				81.83	79.6	681.8	4.981
200411031157	3	450	25%	90.5	0.3733	0.613	9.062		622.3	0.043	487.3				92.46	76.9	696.8	5.002
200411031323	3	450	25%	90.7	0.3448	0.645	9.101		624.2	0.043	488.9				85.12	80.3	700.2	5.021
200411040805	3	450	25%	89.5	0.2710	0.560	8.764		619.6	0.044	503.9				64.56	70.1	653.0	4.828
200411040944	3	450	25%	90.4	0.1699	0.636	8.766		610.5	0.044	503.1				40.61	78.9	658.6	4.795
200411041120	3	450	25%	90.4	0.2284	0.683	8.642		607.5	0.044	499.0				54.69	83.5	653.6	4.803
			ave	90.4	0.2860	0.629	8.867	#DIV/0!	619.5	0.044	496.8	#DIV/0!	#DIV/0!	#	69.67	78.3	672.4	4.910
			stdev	0.4	0.0710	0.038	0.171	#DIV/0!	7.6	0.001	6.7	#DIV/0!	#DIV/0!	#	18.27	4.1	20.3	0.099
			COV	0.5%	24.8%	6.0%	1.9%	#DIV/0!	1.2%	0.014	1.4%	#DIV/0!	#DIV/0!		26.2%	5.3%	3.0%	2.0%
200411030833	4	450	5%	18.40	1.7260	4.613	13.905		1516.7	0.231	421.1				100.91	135.8	252.9	2.888
200411031011	4	450	5%	18.40	1.7208	4.441	14.087		1508.7	0.224	415.9				102.15	131.2	257.8	2.890
200411031157	4	450	5%	18.40	1.8648	4.355	14.135		1495.8	0.217	412.1				110.04	131.0	260.5	2.884
200411031323	4	450	5%	18.40	1.7624	4.366	14.115		1500.8	0.213	412.1				104.72	129.7	260.0	2.893
200411040805	4	450	5%	18.37	1.3652	4.161	13.958		1488.9	0.224	424.5				79.75	126.7	252.5	2.814
200411040944	$\overline{4}$	450	$5\%$	18.37	1.2053	4.197	13.806		1470.2 0.211		422.7				69.57	126.9	250.3	2.788
200411041120	4	450	5%	18.40	1.2490	4.375	13.564		1466.2 0.211		420.2				73.47	129.6	247.7	2.799
			ave	18.4	1.5562	4.358	13.938	#DIV/0!	1492.5	0.219	418.4	#DIV/0!	#DIV/0!	#	91.51	130.1	254.5	2.851
			stdev	0.0	0.2732	0.151	0.205	#DIV/0!	18.9	0.008	5.0	#DIV/0!	#DIV/0!		16.66	3.1	5.0	0.048
			COV	0.1%	17.6%	3.5%	1.5%	#DIV/0!	1.3%	0.035	1.2%	#DIV/0!	#DIV/0!	#	18.2%	2.4%	2.0%	1.7%

**BUG Test Results for PEMS#1 in grams per hp- hour** 











# **BUG Test Results for PEMS#3 in grams per hour**



## **BUG Test Results for PEMS#3 in grams per hp- hour**







#### BUG Test Results for PEMS #4 in grams per hp- hour

	$THC$ (g/hr)				CO (g/hr)				$NO_x$ (g/hr)				$CO2$ (g/hr)	PM(ghr)			
												IPEMS1 PEMS2 PEMS4IPEMS1 PEMS2 PEMS3 PEMS4IPEMS1 PEMS2 PEMS3 PEMS4IPEMS1 PEMS2 PEMS3 PEMS4IPEMS1 PEMS3					
100% load																	
Paired T-Test	0.000	0.002	0.000	0.073	0.000	0.000	0.000	0.000	0.000	0.018	0.002	0.196	0.141	0.505	0.000	0.000	0.000
Unpaired T-test	0.000	0.000	0.000	0.555	0.004	0.003	0.000	0.000	0.000	0.018	0.001	0.164	0.1 14	0.493	0.000	0.000	0.000
F-Test	0.000	0.000	0.000	0.638	0.725	0.925	0.000	0.043	0.894	0.002	0.630	0.001	0.001	0.000	0.938	0.000	0.750
65% load																	
Paired T-Test	0.001	0.000	0.001	0.709	0.000	0.000	0.002	0.000	0.000	0.290	0.000	0.002	0.000	0.001	0.000	0.000	0.000
Unpaired T-test	0.000	0.000	0.000	0.892	0.000	0.000	0.000	0.000	0.000	0.304	0.000	0.001	0.000	0.000	0.000	0.000	0.000
lF-Test	0.000	0.002	0.000	0.439	0.417	0.715	0.000	0.002	0.106	0.001	0.360	0.014	0.003	0.001	0.624	0.000	0.325
25% load																	
<b>Paired T-Test</b>	0.000	0.000	0.001	ΝA	ΝA	ΝA	0.012	0.000	0.000	0.001	0.000	0.001	0.001	0.000	0.000	0.000	0.001
<b>Unpaired T-test</b>	0.000	0.000	0.000	ΝA	ΝA	ΝA	0.004	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000	0.000
lF-Test	0.000	0.085	0.000	NA	ΝA	ΝA	0.000	0.001	0.342	0.000	0.934	0.007	0.002	0.000	0.036	0.000	0.061
5% load																	
<b>Paired T-Test</b>	0.001	0.010	0.004	0.045	0.002	ΝA	0.048	0.000	0.283	NA	0.000	0.003	0.002	ΝA	0.823	0.000	<b>NA</b>
Unpaired T-test	0.000	0.173	0.001	0.269	0.000	0.000	0.035	0.000	0.398	0.000	0.000	0.000	0.000	0.000	0.856	0.000	0.000
F-Test	0.001	0.399	0.000	0.511	0.044	0.054	0.000	0.199	0.374	0.001	0.173	0.009	0.000	0.000	0.056	0.000	0.450

**Appendix E – Statistical Comparisons between FRM and PEMS for BUG Testing (in g/hr)** 



#### **Appendix F – Chassis Dynamometer Integrated Individual Test Results for Each Measurement Device**

Integrated Test Results for the FRM in grams per cycle



## **Integrated Test Results for PEMS#1 in gram per cycle**

Date							Mass Emission g/cycle			<b>Exh Flow</b>			Concentration Measured (wet)		
yyymmddhhmm	Trace	Cycle Dur Load Hp		<b>THC</b>	CO	kNOx	NO <sub>2</sub>	CO <sub>2</sub>	<b>PM</b>	scfm			THC ppm CO ppm NOx ppm CO2%		kH
200504211014	<b>UDDS</b>	1062	76.2	1.267	35.3	64.1	0.92	13999.5		361.1	10.3	120	166.2	2.97	0.918
200504211054	<b>UDDS</b>	1063	77.8	1.486	28.3	65.3	1.29	14152.7		359.7	12.7	94	170.4	2.99	0.913
200504211152	<b>UDDS</b>	1063	79.3	0.966	26.9	66.3	1.51	14430.6		364.2	7.6	86	171.3	2.99	0.906
200504211413	<b>UDDS</b>	1022	81.2	1.391	28.9	64.0	0.06	14330.2		368.8	11.7	98	163.0	3.07	0.929
200504211446	<b>UDDS</b>	1063	80.6	1.097	29.8	68.3	0.30	14793.6		367.5	8.7	98	169.2	3.06	0.924
200504220900	<b>UDDS</b>	1063	78.1	1.182	37.2	65.3	1.78	14457.3		375.4	9.0	116	154.0	2.93	0.963
200504220935	<b>UDDS</b>	1063	77.3	1.207	34.7	64.3	0.02	14286.9		364.5	9.7	111	155.3	2.97	0.959
		ave	78.6	1.23	31.6	65.4	0.8	14350	#DIV/0!	366	9.97	103.3		2.998	0.930
		stdev	1.8	0.17	4.0	1.5	0.7	252.0	#DIV/0!	5.3	1.76	12.4		0.051	0.022
		COV	2.3%	14.2%	0.1	2.3%	85.9%	1.8%	#DIV/0!	1.4%	17.7%	12.0%		1.7%	2.4%
200504201057	50CRUISE	760	203.0	1.086	27.7	99.7	1.58	24985.4		678.7	7.5	$\overline{87}$	209.7	4.84	0.929
200504201233 50CRUISE		760	203.1	0.870	29.3	102.0	2.10	24914.6		678.2	6.3	92	214.6	4.83	0.929
200504201315 50CRUISE		760	200.9	0.827	29.0	102.0	2.55	24629.6		672.1	5.5	93	216.0	4.81	0.930
200504201405	50CRUISE	694	222.3	0.674	27.9	102.7	2.54	24796.4		727.9	4.7	93	223.9	5.11	0.925
200504201443	50CRUISE	760	203.1	1.296	25.2	102.5	1.23	25431.1		683.1	9.0	80	212.7	4.89	0.930
200504201514 50CRUISE		760	206.0	1.218	25.8	103.6	1.57	25536.8		682.2	8.4	81	216.2	4.92	0.927
200504210814 50CRUISE		760	200.9	1.266	35.4	97.8	2.17	24997.6		684.3	8.8	111	203.1	4.82	0.931
		ave	205.6	1.03	28.6	101.5	2.0	25042	#DIV/0!	686.7	7.16	91.0		4.890	0.929
		stdev	7.6	0.24	3.3	2.0	0.5	329	#DIV/0!	18.6	1.68	10.4		0.106	0.002
		COV	3.7%	23.6%	11.7%	2.0%	26.1%	1.3%	#DIV/0!	2.7%	23.5%	11.4%		2.2%	0.2%
		missed test start on 20054201405 and 200504191505													
200504191000 NTE 1290		1112	209.1	1.181	31.3	183.1	5.32	33293.5		537.8	7.8	92	369.1	6.07	0.936
200504210849 NTE 1290		1113	202.9	1.588	37.5	165.9	5.29	32331.3		523.7	11.0	118	336.5	6.05	0.927
200504191046 NTE 1500		1112	233.0	0.930	39.5	193.1	6.37	37841.0		647.9	5.0	97	325.2	5.75	0.936
200504220740 NTE 1500		1113	243.0	1.233	38.3	203.2	8.26	39287.1		653.4	6.6	93	344.3	5.92	0.920
200504191234 NTE 1770		1113	268.6	2.314	46.5	212.5	4.26	47913.3		822.1	9.6	91	275.8	5.77	0.934
200504220814 NTE_1770		1113	277.0	2.234	46.3	214.5	5.41	48301.1		814.1	9.3	91	284.1	5.89	0.924
200504191326 VTE_Stepped		1533	220.6	1.952	50.1	222.8	5.18	50357.8		608.3	7.8	89	256.0	5.44	0.933
200504191505 VTE_Stepped		1419	224.5	1.683	44.9	203.7	6.74	47711.3		621.4	7.0	84	249.6	5.42	0.918
200504191627 VTE_Stepped		1533	223.7	2.173	40.0	215.2	7.17	50900.1		611.8	8.5	69	248.6	5.49	0.914
200504210927 VTE_Stepped		1532	224.9	1.019	51.4	219.5	4.82	51343.0		614.9	4.0	93	259.0	5.60	0.920
200504211237 VTE_Stepped		1533	231.9	1.153	34.8	233.4	8.61	51980.3		608.3	4.3	60	270.7	5.62	0.921
200504211320 VTE Stepped		1533	233.7	1.850	36.5	231.2	6.42	53003.7		617.3	7.0	63	265.9	5.74	0.930

 **Integrated Test Results for PEMS#2 in gram per cycle** 







## Integrated Test Results for PEMS #4 in grams per cycle

					CO				NO <sub>x</sub>					$\overline{2}$		
									PEMS1 PEMS2 PEMS4 PEMS1 PEMS2 PEMS3 PEMS4 PEMS1 PEMS2 PEMS3 PEMS4 PEMS1 PEMS2 PEMS3 PEMS4 PEMS3							
<b>UDDS</b>																
<b>Paired T-Test</b>	0.114	0.157	0.011	0.132	0.000	0.000	NA	0.415	0.000	0.000	0.000	0.147	0.005	0.000	0.000	0.000
<b>Unpaired T-test</b>	0.008	0.074	0.016	0.006	0.016	0.002	NA	0.010	0.000	0.000	0.000	0.000	0.157	0.000	0.000	0.000
F-Test	0.005	0.110	0.027	0.640	0.527	0.466	<b>NA</b>	0.024	0.378	0.761	0.805	0.016	0.464	0.729		$0.092$ 0.974
50 mph <del>cruise</del>														PM		
Paired T-Test	0.017	0.605	<b>NA</b>	0.266	0.001	0.001	NA	0.510	0.001	0.000	<b>NA</b>	0.144	0.016	0.000	NA	0.138
<b>Unpaired T-test</b>	0.001	0.653	<b>NA</b>	0.676	0.012	0.014	<b>NA</b>	0.010	0.006	0.000	<b>NA</b>	0.000	0.025	0.000	<b>NA</b>	0.094
F-Test	0.457	0.901	<b>NA</b>	0.115	0.434	0.196	<b>NA</b>	0.203	0.212	0.737	<b>NA</b>	0.023	0.821	0.065	<b>NA</b>	0.018
<b>NTE 1290</b>										CO						
Paired T-Test	NA	0.351	<b>NA</b>	<b>NA</b>	0.055	0.015	<b>NA</b>	NA	0.162	0.117	<b>NA</b>	<b>NA</b>	0.174	0.062	<b>NA</b>	0.699
Unpaired T-test	<b>NA</b>	0.156	<b>NA</b>	ΝA	0.090	0.111	NA	<b>NA</b>	0.434	0.219	<b>NA</b>	<b>NA</b>	0.261	0.063	NA	0.923
F-Test	<b>NA</b>	0.901	<b>NA</b>	<b>NA</b>	0.758	0.941	<b>NA</b>	<b>NA</b>	0.772	0.702	<b>NA</b>	<b>NA</b>	0.601	0.665	<b>NA</b>	0.807
<b>NTE 1550</b>																
Paired T-Test	<b>NA</b>	0.380	<b>NA</b>	<b>NA</b>	0.016	0.029	NA		0.030	0.008	<b>NA</b>	<b>NA</b>	0.253	0.083	<b>NA</b>	NA
Unpaired T-test	<b>NA</b>	0.183	<b>NA</b>	<b>NA</b>	0.005	0.009	<b>NA</b>	<b>NA</b>	0.095	0.050	<b>NA</b>	<b>NA</b>	0.409	0.009	<b>NA</b>	NA
<b>F-Test</b>	<b>NA</b>	0.653	<b>NA</b>	NA	0.659	0.567	<sub>NA</sub> NA	<b>NA</b>	0.872	0.950	<b>NA</b>	NA	0.601	0.788	<b>NA</b>	NA
<b>NTE 1770</b>																
Paired T-Test	NA	0.574	<b>NA</b>	<b>NA</b>	0.055	0.065	NA		0.067	0.082	<b>NA</b>	<b>NA</b>	0.130	0.008	<b>NA</b>	<b>NA</b>
<b>Unpaired T-test</b>	<b>NA</b>	0.426	<b>NA</b>	<b>NA</b>	0.006	0.014	<b>NA</b>	<b>NA</b>	0.009	0.015	<b>NA</b>	<b>NA</b>	0.158	0.011	<b>NA</b>	NA
F-Test	NA	0.383	<b>NA</b>	<b>NA</b>	0.139	0.178	NANA	<b>NA</b>	0.187	0.054	<b>NA</b>	NA	0.580	0.896	<b>NA</b>	<b>NA</b>
<b>NTE Stepped</b>																
Paired T-Test	NA	0.083	0.784	NA	0.005	0.001	<b>NA</b>	NA	0.001	0.005	0.136	NA	0.003	0.001	0.001	0.011
<b>Unpaired T-test</b>	<b>NA</b>	0.171	0.209	<b>NA</b>	0.004	0.001	<b>NA</b>	NA	0.002	0.000	0.021	<b>NA</b>	0.043	0.000	0.000	0.103
F-Test	<b>NA</b>	0.472	0.032	NA	0.030	0.133	<b>NA</b>	<b>NA</b>	0.532	0.104	0.002	<b>NA</b>	0.757	0.188	0.228	0.979

**Appendix G – Statistical Comparisons between FRM and PEMS for Integrated Chassis Dynamometer Cycles (p-values)** 

# **Appendix H – bs and fs Emissions Results for NTE Steady State Cycles**







# **Appendix I – Statistical Comparisons for NTE Steady State Cycles (g/bhp-hr)**

# **Appendix J Statistical Comparisons for NTE Stepped Cycles (g/bhp-hr)**



	Setup Information		<b>Other Values</b>					PM Values (g/cycle)				Exh Flow		<b>Ambient Conditions</b>	
<b>Test Name</b>	Trace	Cycle Dur Vmix_m3		SecDF	<b>FRM</b>	PEMS <sub>3</sub>	PEMS <sub>5</sub>	PEMS <sub>7</sub>	PEMS 8	FRM EC	FRM OC	scfm		Temp C Baro mmHg	kH
200504211014	<b>UDDS</b>	1060	1000.7	2.516	3.55	3.14	5.65	3.09	5.16	2.45	0.59	$\overline{347}$	19.5	752.7	0.946
200504211054	<b>UDDS</b>	1060	1000.7	2.516	3.61	3.03	6.39	3.08	5.26	2.55	0.63	348	20.2	752.6	0.942
200504211152	<b>UDDS</b>	1060	1000.8	2.516	3.77	3.24	6.77	3.26	5.54	n/a	n/a	353	22.1	752.4	0.929
200504211413	<b>UDDS</b>	1060	1000.8	2.516	3.81	3.24	5.86	3.34	5.41	n/a	n/a	353	24.9	751.7	0.965
200504211446	<b>UDDS</b>	1060	1000.7	2.516	4.00	3.43	6.10	3.48	5.73	n/a	n/a	356	25.1	751.2	0.957
200504220900	<b>UDDS</b>	1060	1000.7	2.516	4.07	3.02	5.60	n/a	5.16	n/a	n/a	362	19.6	751.0	0.978
200504220935	<b>UDDS</b>	1060	1000.6	2.516	3.87	2.87	5.34	n/a	4.99	n/a	n/a	349	18.4	751.1	0.976
		ave	1001	2.516	3.81	3.14	5.96	3.25	5.32	2.50	0.61	353	21.4	751.8	0.956
		stdev	0.054	0.000	0.19	0.19	0.50	0.17	0.25	0.07	0.03	5	2.7	0.8	0.018
		COV	0%	0%	5%	6%	8%	5%	5%	3%	4%	1.5%	12.7%	0.1%	1.9%
200504201057	50CRUISE	757	714.5	2.516	4.30	4.14	6.74	3.53	5.64	3.08	0.85	664	18.1	753.5	0.954
200504201233	50CRUISE	757	714.6	2.516	4.20	3.82	6.45	3.46	5.55	3.03	0.81	664	20.5	752.8	0.956
200504201315	50CRUISE	757	714.3	2.516	3.98	3.95	6.31	3.33	5.32	2.86	0.74	659	20.7	752.8	0.956
200504201405	50CRUISE	757	714.3	2.516	4.16	3.76	6.29	3.38	5.51	2.84	0.77	665	21.1	752.8	0.954
200504201443	50CRUISE	757	714.4	2.516	4.29	3.87	6.30	3.44	5.60	2.97	0.61	665	21.0	752.8	0.959
200504201514	50CRUISE	757	714.3	2.516	4.95	3.66	6.10	partial	5.40	2.74	0.83	667	20.8	752.5	0.959
200504210814	50CRUISE	757	714.5	2.516	3.42	3.80	6.74	3.51	5.60	2.42	0.59	672	17.0	753.4	0.957
		ave	$\overline{714}$	2.516	4.19	3.86	6.42	3.44	5.52	2.85	0.74	665	19.9	752.9	0.957
		stdev	0.098	0.000	0.45	0.15	0.24	0.08	0.12	0.22	0.11	4.1	1.6	0.4	0.002
		COV	0%	0%	11%	4%	4%	2%	2%	8%	14%	0.6%	8.3%	0.0%	0.2%
200504191000	<b>NTE 1290</b>	1110	1047.3	1.573	4.12	4.18	6.02	n/a	n/a	n/a	n/a	516	16.5	750.4	0.963
200504210849	NTE 1290	1110	1047.2	2.516	3.44	3.25	5.52	2.90	4.88	2.42	0.59	508	17.2	753.2	0.955
200504191046	NTE 1500	1110	1047.2	1.573	4.49	6.26	6.43	n/a	n/a	n/a	n/a	629	17.5	750.4	0.958
200504220740	NTE 1500	1110	1047.2	2.516	n/a	4.44			6.52	n/a	n/a	640	18.9	749.5	0.959
200504191234	NTE 1770	1110	1047.1	2.061	7.81	11.67	8.25	invalid	9.69	n/a	n/a	800	19.8	750.2	0.962
200504220814	NTE 1770	1110	1047.1	2.516	n/a	8.28	10.15	n/a	10.53	n/a	n/a	805	19.7	750.0	0.975
200504191326 NTE_Stepped		1530	1443.9	1.139	invalid	4.76	n/a	partial	6.62	n/a	n/a	591	20.0	750.0	0.963
200504191505 NTE Stepped		1530	1443.9	2.516	5.58	n/a	6.23	4.20	7.70	n/a	n/a	599	19.6	750.3	0.951
200504191627 NTE_Stepped		1530	1443.9	2.516	5.38	n/a	6.00	4.18	7.41	n/a	n/a	603	19.2	750.4	0.948
200504210927 NTE_Stepped		1530	1444.1	2.516	5.75	4.84	7.42	4.40	8.33	3.61	1.27	601	18.3	753.1	0.946
200504211237 NTE Stepped		1530	1443.9	2.516	6.59	5.15	8.65	5.28	10.42	n/a	n/a	596	24.8	752.3	0.949
200504211320 NTE Stepped		1530	2383.0	2.516	7.52	6.44	7.11	6.01	10.35	n/a	n/a	617	25.0	750.9	0.966

**Appendix K – PM data for PEMS 3, 5, 7, 8 and FRM including EC/OC** 



## **Appendix L – Statistical results for PEMS 3, 5, 7, 8, and EC compared to FRM gravimetric**

#### **References**

 1 One Example is Environmental Protection Agency, *Draft Technical Support Document: In-Use Testing for Heavy-Duty Diesel Engines and Vehicles*, EPA Document # 420-D-04-003, June 2004

2 Cocker, D.R. III, Shah, S.D., Johnson, K., Miller, J.W., and Norbeck, J.M., 2004. "Development and Application of a Mobile Laboratory for Measuring Emissions from Diesel Engines. 1. Regulated Gaseous Emissions" Environ. Sci. Technol., Vol. 38, p. 2182-2189.

3 US Environmental Protection Agency, *Test Plan to Determine PEMS Measurement Allowances for the Gaseous Emissions Under the Manufacturer-Run Heavy-duty Diesel Engine In-Use Testing Program*, EPA Docket #OAR-2004-0072-0069, May 2005, p. 9.

4 Kean, A. J., Sawyer, R. J., Harley, R. A., *A Fuel-Based Assessment of Off-Road Diesel Engine Emissions,* J. Air & Waste Manage. Assoc. **50**:1929-1939 (2000)

 $5$  U.S. Patent No. 6,062,092. "System for Extracting Samples from a Stream", May 16, 2000.